



1 The influence of Holocene vegetation changes on topography
2 and erosion rates: A case study at Walnut Gulch Experimental
3 Watershed, Arizona

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10

11 **Abstract**

12 Quantifying how landscapes have responded and will respond to vegetation changes is an
13 essential goal of geomorphology. The Walnut Gulch Experimental Watershed offers a unique
14 opportunity to quantify the impact of vegetation changes on landscape evolution over geologic
15 time scales. The Walnut Gulch Experimental Watershed (WGEW) is dominated by grasslands at
16 high elevations and shrublands at low elevations. Paleovegetation data suggest that portions of
17 WGEW higher than approximately 1430 m a.s.l. have been grasslands and/or woodlands
18 throughout the late Quaternary, while elevations lower than 1430 m a.s.l. changed from a
19 grassland/woodland to a shrubland c. 2-4 ka. Elevations below 1430 m a.s.l. have decadal time-
20 scale erosion rates approximately ten times higher, drainage densities approximately three times
21 higher, and hillslope-scale relief approximately three times lower than elevations above 1430 m.
22 We leverage the abundant geomorphic data collected at WGEW over the past several decades to



23 calibrate a mathematical model that predicts the equilibrium drainage density in shrublands and
24 grasslands/woodlands at WGEW. We use this model to test the hypothesis that the difference in
25 drainage density between the shrublands and grassland/woodlands at WGEW is partly the result
26 of a late Holocene vegetation change in the lower elevations of WGEW, using the upper
27 elevations as a control. Model predictions for the increase in drainage density associated with the
28 shift from grasslands/woodlands to shrublands are consistent with measured values. Using
29 modern erosion rates and the magnitude of relief reduction associated with the transition from
30 grasslands/woodlands to shrublands, we estimate the timing of the grassland-to-shrubland
31 transition in the lower elevations of WGEW to be approximately 3 ka, i.e., broadly consistent
32 with paleovegetation studies. Our results provide support for the hypothesis that common
33 vegetation changes in semi-arid environments (e.g. from grassland to shrubland) can change
34 erosion rates by more than an order of magnitude, with important consequences for landscape
35 morphology.

36 *Keywords: landscape evolution, drainage density, vegetation cover, Walnut Gulch Experimental*
37 *Watershed*

38

39 **1. Introduction**

40 **1.1. Problem statement**

41 Understanding how climate change controls landscape evolution is a central problem in
42 geomorphology. Climate changes are multifaceted, with changes in temperature (mean and
43 variability), precipitation (mean and variability) and vegetation cover (type and density) often
44 occurring simultaneously. The multifaceted nature of climatic changes can make it difficult to
45 identify which aspects of climate change are most important in driving landscape modification in



46 specific cases. Yet, given the accelerated climatic changes expected to occur in the coming
47 decades, understanding how landforms are likely to respond to specific climatic drivers, acting
48 alone or in concert, is critically important to society (e.g., Pelletier et al., 2015).

49 In the southwestern U.S. the existence of an extensive, regionally correlative fan and
50 valley floor depositional unit (the Q3a unit of Bull, 1991) suggests that the Pleistocene-to-
51 Holocene transition was a major perturbation in what are now semi-arid shrubland-dominated
52 landscapes but which were predominantly pinyon-juniper woodlands at the Last Glacial
53 Maximum (LGM). Where rates of aggradation in modern shrubland drainage basins have been
54 measured, the Pleistocene-to-Holocene transition was associated with more than an order-of-
55 magnitude increase above either LGM or mid-to-late Holocene rates (e.g., Fig. 3.11 of Weldon,
56 1986). The retreat of grasslands and woodlands to higher elevations and their replacement by
57 shrublands likely exposed elevations of the southwestern U.S. between approximately 800 m and
58 at least 1400 m a.s.l. to significant increases in percent bare ground, thus modifying the rainfall-
59 runoff partitioning of hillslopes and their resistance to fluvial/slope-wash erosion. In especially
60 arid areas of the southwestern U.S. such as the central Mojave Desert, the range of elevations
61 affected by late Quaternary conversions of grasslands/woodlands to shrublands extends to
62 elevations as high as 1800 m a.s.l. (Pelletier, 2014).

63 Given the strong correlation between percent bare ground and drainage density in the
64 southwestern U.S. (Melton, 1957), it has been hypothesized that modern shrublands sufficiently
65 high in elevation to have been grasslands or woodlands at LGM underwent large increases in
66 drainage density during the Pleistocene-to-Holocene transition. Such an expansion of the fluvial
67 drainage network could have mobilized hillslope deposits stored as colluvium during the last
68 glacial epoch, mobilizing large volumes of sediment into the fluvial system during the transition



69 to the present interglacial (Bull, 1991; Pelletier, 2014). This hypothesis is consistent with
70 measured ages of the onset of aggradation in valley floor and alluvial fan depositional zones in
71 the central Mojave Desert, in which aggradation occurs earliest (c. 15 ka) in depositional zones
72 fed by source regions with relatively low elevations in the 800-1800 m a.s.l. range and
73 progressively later in areas fed by eroding regions at higher elevations (Pelletier, 2014).
74 Alternative explanations for the punctuated nature of late Quaternary aggradation in the
75 southwestern U.S. invoke changes in the frequency and/or magnitude of floods (Antinao and
76 McDonald, 2013b) and argue that hillslope vegetation changes played a limited role (Antinao
77 and McDonald, 2013a). To date, tests of the paleovegetation change hypothesis in the
78 southwestern U.S. have been limited to studies of the timing of deposition on valley-floor
79 channels and alluvial fans. Erosion of the source drainage basins themselves has been relatively
80 understudied.

81 The Walnut Gulch Experimental Watershed (WGEW) provides an excellent opportunity
82 to test the paleovegetation change hypothesis in a drainage basin that has been extensively
83 monitored for decades. The western, low-elevation portion of WGEW is currently a shrubland
84 while the eastern portion is predominantly a grassland (a small area of woodland occupies the
85 highest elevations). Paleovegetation data, however, indicate that all of WGEW was a grassland
86 or woodland until just a few thousand years ago. Scientists working at WGEW have measured
87 rates of sediment export from watersheds using sediment samplers (e.g., Nichols et al., 2008) and
88 rates of sediment redistribution within watersheds using anthropogenic radionuclides (e.g., ¹³⁷Cs)
89 and rare-earth element tracers (e.g., Nearing et al., 2005; Polyakov et al., 2009). These data, in
90 addition to the results of analyses of airborne lidar data presented here, make it possible to
91 calibrate every parameter of a mathematical model that predicts the equilibrium drainage density



92 of the landscape under different dominant vegetation types. In this paper we use this
93 mathematical model to test the hypothesis that grassland/woodland-to-shrubland vegetation
94 changes in the lower elevations of WGEW drove large increases in drainage density and erosion
95 rates and a decrease in hillslope-scale relief.

96 **1.2. Study Site**

97 *1.2.1. Geology and Soils*

98 The Walnut Gulch Experimental Watershed (WGEW) is part of the U.S. Department of
99 Agriculture (USDA) Agricultural Research Service's (ARS) Southwest Watershed Research
100 Center (SWRC). WGEW is located at the boundary between the Chihuahuan and Sonoran
101 Deserts and elevations of between approximately 1300 and 1600 m a.s.l. The approximately 150
102 km² watershed has a planar geometry at large spatial scales, dipping to the WSW with a slope of
103 approximately 1.5%, i.e., approximately 230 m of elevation change over a distance of 15 km.

104 The bedrock geology of WGEW includes sedimentary, plutonic, and volcanic rocks of
105 Precambrian to late Cenozoic age (Fig. 1). Due to the complex nature of the rock types exposed
106 in the southern portion of the watershed, we focus this study on the northern portion of the
107 watershed, which is dominated by the Gleeson Road Conglomerate (GRC).

108 The GRC was derived primarily from erosion of the Dragoon Mountains to the east of
109 WGEW and is estimated to be Plio-Pleistocene in age by Osterkamp (2008), who noted: "The
110 upper part of the Gleeson Road Conglomerate is probably equivalent both stratigraphically and
111 in age to the Plio-Pleistocene upper basin fill of Brown et al. (1966). To the west and northwest,
112 along the axis of the San Pedro River Valley, the upper part of the Gleeson Road Conglomerate
113 grades into fine-grained fluvial and lacustrine beds of the Plio-Pleistocene St. David
114 Formation (Gray 1965, Melton 1965)." The GRC dips gently (5-8°) to the north and northwest



115 (Gilluly, 1956, plate 5), a fact which may contribute to the generally longer and more gently
116 sloping south-facing hillslopes relative to north-facing hillslopes in the watershed. The tilted
117 strata of the GRC were beveled to a gently sloping topographic surface and incised into during
118 Quaternary time. Incision of the GRC was driven by uplift/tilting of the fan and/or by incision of
119 the San Pedro River, triggering headward erosion of Walnut Gulch and its tributaries.

120 The northern portion of WGEW is composed of the Whetstone Pediment of Bryan (1926)
121 and is divided into two parts: the “dissected Whetstone Pediment” (DWP) at elevations below
122 approximately 1430 m a.s.l. and the “upper Whetstone Pediment” (UWP) at higher elevations
123 (Fig. 1). The DWP is distinguished from the UWP by its lower relief and less well-developed
124 soils. Differences between the DWP and UWP have been attributed primarily to headward
125 extension of tributaries resulting from late Quaternary incision of the San Pedro River (Cooley,
126 1968) as well as renewed river and tributary incision following livestock grazing beginning c.
127 1880 (Renard et al., 1993; Osterkamp, 2008). However, the boundary between the DWP and
128 UWP coincides with the transition from the modern shrubland to the grassland (Fig. 2). As such,
129 we hypothesize that vegetation cover and its changes over geologic time scales have also
130 contributed to differences between the DWP and UWP.

131 Deep sandy gravel loams of the Blacktail-Elgin-Stronghold-McAllister-Bernardino
132 Group occur in areas of the UWP. In the lower, western part of the watershed, soils are in the
133 Luckyhills-McNeal Group. Soils of the Luckyhills-McNeal Group tend to be sandy and gravelly
134 loams that are immature compared with soils of the Blacktail-Elgin-Stronghold-McAllister-
135 Bernardino Group. The A horizon of the Luckyhills-McNeal Group is typically absent, having
136 been removed by late Quaternary erosion (Breckenfeld, 1994). The boundary between these two



137 soil groups coincides with the boundary between the DWP and UWP and the transition from the
138 modern shrubland to the grassland.

139 *1.2.3. Climate and Vegetation*

140 Mean annual temperature at Tombstone (located in the western portion of WGEW at an
141 elevation of 1384 m a.s.l.) is 17.6°C and mean annual precipitation is approximately 300 mm.

142 There is both a winter and summer rainy season, but approximately 70% of rainfall occurs in the
143 summer months and the greater intensity of summer storms means that runoff results almost
144 exclusively during the summer season of July through September. Mean annual precipitation is
145 approximately 10% higher at the highest elevations of the watershed relative to the lowest
146 elevations (Nearing et al., 2015).

147 Modern vegetation cover in the U.S.-Mexico Borderlands region closely follows
148 elevation, with desert scrub occurring below approximately 1500 m, grasslands from 1400–
149 1700 m, lower encinal (“encina” is Spanish for “oak”) from 1700–2600 m, upper encinal from
150 1900–2600 m, and forest from 2200–2600 m (Wagner, 1977). For each zone, the highest
151 elevations are reached on dry, south- and west-facing slopes and the lowest elevations on north-
152 facing slopes and valley bottoms.

153 Paleovegetation records from the Borderlands region suggest that the western portions of
154 WGEW have transitioned from a grasslands/woodland to a shrubland over the past few thousand
155 years, while the eastern half of the watershed has been a grassland/woodland for at least the past
156 40,000 yr and likely since the penultimate interglacial period. Low stalagmite ¹⁸O values at the
157 Last Glacial Maximum (LGM) in the Cave of the Bells paleoclimate record indicate that
158 conditions were much wetter and cooler during LGM (Wagner et al., 2010), in agreement with
159 paleovegetation studies (Betancourt et al., 1990; Anderson, 1993; Betancourt et al., 2001;



160 Arundel, 2002; Holmgren et al., 2003; Holmgren, 2005). Packrat midden records indicate the
161 presence of grasslands and/or pinyon–juniper woodlands at LGM in what is currently
162 Chihuahuan desert scrub at elevations of 1200–1400 m a.s.l. (Betancourt et al., 2001).
163 Holmgren (2005) documented the presence of the primary grass species at WGEW (*Bouteloua*
164 *eriopoda*) (currently abundant only at elevations above approximately 1430 m a.s.l.) at an
165 elevation of 1287 m c. 4750 ^{14}C yr B.P. Elements of the modern shrubland, such as *Larrea*
166 *tridentate*, appeared as late as 2190 ^{14}C yr B.P. at 1287 m a.s.l. based on macrofossils, but may have
167 been present as early as 4095 ^{14}C yr B.P. based on pollen. As such, the latest Quaternary transition
168 from grasslands/woodlands to shrublands in WGEW occurred gradually and was completed only
169 within the past few thousand years.

170 **1.2.4. Intensive monitoring sites**

171 WGEW is home to two intensive monitoring sites, one in the shrubland of the DWP and
172 the other in the grassland of the UWP. These sites provide the data necessary, in conjunction
173 with the topographic analyses presented here, to calibrate a mathematical model that predicts the
174 equilibrium drainage density as a function of vegetation cover and to test the hypothesis that
175 differences in landscape morphology and erosion rates between the northwestern and
176 northeastern portions of WGEW are partly the result of a transition from grassland/woodland to
177 shrubland within the past few thousand years in the northwestern portion of the drainage basin
178 that did not occur in the northeastern portion.

179 Watersheds 63.102, 63.103, 63.104, 63.105, and 63.106 are located in shrublands at an
180 elevation of approximately 1370 m a.s.l. in what is referred to as “Lucky Hills,” which has been
181 the site of a variety of intensive scientific studies since the 1960s. Cover during the rainy season
182 at Lucky Hills is approximately 25% bare soil, 25% canopy, and 50% erosion pavement (rocks).



183 Dominant vegetation includes: Creosote (*Larrea tridentata*, shrub) and Whitethorn (*Acacia*
184 *constricta*, shrub), with lesser populations of Desert Zinnia (*Zinnia acerosa*, shrub), Tarbush
185 (*Flourensia cernua*, shrub), and sparse Black Grama (*Bouteloua eriopoda*, grass). The matrix
186 material of surface layer is composed of 60% sand, 25% silt, and 15% clay. Sediment from the
187 watershed is monitored with a supercritical flume with an automatic traversing slot sampler
188 (Simanton et al., 1993).

189 Watershed 63.112 is located in the Kendall grassland site at WGEW, approximately 10
190 km east of Lucky Hills and at an elevation of approximately 1525 m. The site is predominantly
191 covered by grass and forbs with some shrubs and succulents with a combined canopy cover of
192 approximately 35%. Ground cover during the rainy season has been measured at 28% rock, 42%
193 litter, and 14% plant basal cover (Nearing et al., 2007). Compared to the 25% bare soil at Lucky
194 Hills, the bare soil exposed at Kendall is negligible (i.e. a few percent or less). Historically, the
195 dominant desert grassland bunchgrasses at the site have been black grama (*Bouteloua eriopoda*),
196 side-oats grama (*B. curtipendula*), three-awn (*Aristida sp.*), and cane beardgrass (*Bothriochloa*
197 *barbinodis*) (King et al., 2008), and more recently, Lehmann lovegrass (*Eragrostis lehmanniana*)
198 (Moran et al., 2009).

199 Nearing et al. (2005) used spatially distributed ^{137}Cs measurements to quantify
200 fluvial/slope-wash erosion and deposition rates within and from watersheds 63.103 (Lucky Hills)
201 and 63.112 (Kendall). Nearing et al. (2005) found that in Lucky Hills the fraction of the drainage
202 basin experiencing erosion was much higher (85%) than in Kendall (53%), where erosion and
203 deposition rates were lower and approximately balanced such that the rate of net sediment
204 exported from the Kendall watershed was more than a factor of ten lower than from the Lucky
205 Hills watershed. These observations are consistent with sediment yields measured from 1995-



206 2005 that are more than ten times higher in drainage basins of similar size at Lucky Hills (231 t
207 $\text{km}^{-2} \text{yr}^{-1}$ in watershed 102 (area of 1.46 Ha)) than at Kendall (7 t $\text{km}^{-2} \text{yr}^{-1}$ in watershed 112
208 (area of 1.86 Ha)) (Nearing et al., 2007). Assuming a soil bulk density of 1500 kg m^{-3} , these
209 sediment yields correspond to mean erosion rates of approximately $1.5 \times 10^{-4} \text{ m yr}^{-1}$ in the
210 shrublands of Lucky Hills and $5 \times 10^{-6} \text{ m yr}^{-1}$ in the grasslands of Kendall (Table 1). Hillslopes are
211 slightly steeper in Kendall, so if anything we would expect erosion rates to be higher at Kendall
212 than at Lucky Hills if slope gradient were the dominant factor in controlling erosion rates.
213 Nearing et al. (2005) interpreted the differences in erosion rates between Lucky Hills and
214 Kendall to be primarily a function of vegetation cover, i.e. “hydrologic response differences as a
215 function of vegetation differences are probably largely responsible for the differences in hillslope
216 erosion rates between the two watersheds. If flows are more concentrated and vegetative cover is
217 less, as on the Lucky Hills site, flow shear stresses and stream power will tend to be greater,
218 resulting in a greater hydrologic potential for erosion. Also important is probably the higher litter
219 cover and plant basal area cover on the grassland site that would have a direct protective effect
220 against erosion.” This interpretation is consistent with the conclusions of Abrahams et al. (1995)
221 and Parsons et al. (1996), who emphasized the role of vegetation cover in controlling
222 fluvial/slope-wash erosion rates in their plot-scale studies at WGEW. In this paper we explore
223 the implications of these erosion rate differences for landscape morphology and topographic
224 evolution of WGEW over geologic time scales.

225

226 **2. Methods**

227 **2.1. Topographic analysis**



228 In this section we describe the methods used to quantify the similarities and differences in
229 landscape morphology between the modern grassland and shrubland sites, with an eye toward
230 providing the data necessary to calibrate the mathematical model described in section 2.2. In all
231 of the topographic analyses described in this section we used a 1 m pixel⁻¹ Digital Elevation
232 Model (DEM) for WGEW derived from airborne laser swath mapping (Heilman et al., 2008).
233 Prior to the analyses, the DEM was smoothed with an Optimal Weiner Filter (OWF), following
234 the approach of Pelletier (2013), to remove small-scale variability related to errors related to
235 distinguishing ground from vegetation points and other imperfections of the lidar-derived DEM.
236 The smoothing did not significantly alter slope gradients but did significantly reduce anomalous
237 curvature values related to DEM imperfections.

238 The drainage density in the shrubland areas appears to be much higher than in the
239 grassland areas (Fig. 3). To quantify this difference we used the method developed by Pelletier
240 (2013) to identify the network of valley bottoms (i.e. where water flow is localized and fluvial
241 processes are responsible for the majority of erosion) in the vicinity of Lucky Hills and Kendall.
242 In this method, the DEM is filtered using the OWF, the contour curvature is computed at every
243 pixel, and valley heads are identified as the areas closest to the divides where the contour
244 curvature exceeds a user-defined threshold value. In Pelletier (2013) and in this paper a threshold
245 contour curvature of 0.1 m⁻¹ was used for valley head identification (i.e., the method of Pelletier
246 (2013) was used without modification or parameter tuning). Once valley heads are identified, a
247 multiple-flow-direction routing algorithm is used to route a unit of runoff from each valley head
248 to identify the valley bottoms downstream. In this study we used the distance along flow lines
249 from topographic divides to the first valley bottom as a measure of the extent of the drainage
250 network. We computed the mean value of this hillslope length for all pixels entering valley



251 heads. This mean value can be compared to the prediction of a mathematical model that
252 computes the mean distance along flow lines from topographic divides to valley heads as a
253 function of colluvial and fluvial transport coefficients. Hillslope length measured in this way is
254 inversely related to drainage density (Horton, 1945). Its mean value contains information
255 equivalent to drainage density, but it has the advantage that it is a mappable quantity (Tucker et
256 al., 2001). The drainage network analysis was performed on representative 1.6 x 1.6 km areas in
257 the vicinity of Lucky Hills and Kendall to quantify the difference in drainage density between
258 shrubland and grassland areas within WGEW.

259 Relief was mapped as the difference between the highest elevation upstream from each
260 pixel along flow lines. The results of the drainage network identification procedure described
261 above were used to limit this analysis to the hillslope pixels only. We then computed the mean
262 hillslope relief within 10 m elevation bins from 1320 m to 1550 m a.s.l. The resulting graph
263 quantifies how the mean hillslope relief varies with elevation across the shrubland-to-grassland
264 transition.

265 We also computed the mean value of the along-channel slope gradient and curvature (i.e.
266 the Laplacian) as functions of contributing area. Differences in mean curvature as a function of
267 contributing area provide a quantitative signature of how late Holocene vegetation changes have
268 modified the landscape morphology in the vicinity of the hillslope-to-valley-bottom transition.

269 **2.2. Mathematical modeling**

270 In this section we describe the mathematical model used to quantify erosion over
271 geologic time scales and its dependence on landscape morphology at WGEW. The mathematical
272 model is used to predict the equilibrium drainage density, quantified as the mean distance along
273 flow lines from divides to valley bottoms, in both shrublands and grasslands, in order to test the



274 hypothesis that the difference in drainage density observed between the shrublands and
275 grasslands at WGEW can be attributed, in part, to late Holocene vegetation changes in the
276 shrubland portion of the watershed.

277 Erosion at WGEW over geologic time scales can be approximated as the sum of erosion
278 due to colluvial and fluvial/slope-wash processes. Sediment transport by colluvial processes
279 leads to a diffusion equation for topography if slope are uniformly soil-mantled and topographic
280 gradients are modest (Culling, 1960):

$$281 \quad E_c = -D\nabla^2 z \quad (1)$$

282 where E_c is the erosion rate by colluvial processes (defined to be positive if the landscape is
283 eroding), D is diffusivity in $\text{m}^2 \text{yr}^{-1}$, and z is elevation in m. Equation (1) assumes that colluvial
284 sediment flux is equal to the product of a coefficient (i.e., the diffusivity D) and the local slope
285 gradient. This assumption is reasonable for WGEW, where hillslopes are uniformly soil mantled
286 and the mean hillslope gradient is 7%.

287 We assume that fluvial/slope-wash processes in WGEW can be approximated as
288 transport limited. We use the term fluvial/slope-wash to refer to all sediment transport by
289 flowing water, wherever it occurs along the continuum from hillslopes (i.e., as sheet and rill
290 flow) to channels (i.e., fluvial erosion). A transport-limited condition applies to landscapes in
291 which most of the sediment is transported as bed-material load and sediment is readily deposited
292 if the shear stress by flowing water declines with increasing distance along flow pathways. In the
293 alternative detachment-limited end-member model of fluvial/slope-wash erosion, the shear stress
294 required to detach sediment is much larger than the shear stress required to transport it, hence
295 sediment redeposition is rare or nonexistent once detachment/entrainment occurs. Pelletier
296 (2012) addressed the geomorphic conditions under which transport-limited versus detachment-



297 limited conditions are likely to occur, taking into account data for the relative proportion of
298 sediment transported as bed-material load versus wash load, among other factors, using WGEW
299 as a case study. He concluded that among these two end-member models, WGEW is most
300 accurately considered to be transport limited.

301 The assumption of transport-limited conditions implies that fluvial/slope-wash erosion,
302 E_f , is equal to the divergence of the fluvial/slope-wash volumetric unit sediment flux, \mathbf{q}_s (the
303 volumetric sediment flux per unit width of water flow):

$$304 \quad E_f = \frac{1}{\varepsilon_0} \nabla \cdot \mathbf{q}_s . \quad (2)$$

305 where ε_0 is the grain packing density (assumed here to be 0.55, e.g., equivalent to a bulk density
306 of 1500 kg m^{-3} for a grain density of 2700 kg m^{-3}). The divergence of the fluvial/slope-wash
307 sediment flux can, within the valley network, be approximated by the derivative of the
308 volumetric unit sediment flux in the along-valley direction, q_s , with respect to the distance
309 downstream along flow lines, x . Adopting this approximation and summing the colluvial and
310 fluvial/slope-wash components, the total erosion rate at any point on the landscape is thus given
311 by

$$312 \quad E = \frac{1}{\varepsilon_0} \frac{\partial q_s}{\partial x} - D \nabla^2 z . \quad (3)$$

313 In this paper we use equation (3) to predict the equilibrium drainage density, quantified
314 as the mean distance from divides to valley bottoms, associated with grasslands and shrublands
315 at WGEW. Specifically, we substitute an empirical power-law relationship for the mean
316 distance, x , along flow lines from divides to valley bottoms versus contributing area, A , into
317 equation (3) and solve for the value of x where the fluvial erosion rate exceeds the colluvial
318 deposition rate by an amount equal to E .



319 Next, we further parameterize the three terms in equation (3) in terms of measured
320 proxies for erosion rate (e.g. decadal-scale sediment fluxes) and topographic parameters. The
321 erosion rate, E , is constrained using the ratio of the volumetric sediment flux, Q_s , reported by
322 Nearing et al. (2007) and contributing area, A :

$$323 \quad E = \frac{1}{\varepsilon_0} \frac{Q_s}{A}. \quad (4)$$

324 The fluvial erosion term can be written in terms of Q_s using

$$325 \quad \frac{\partial q_s}{\partial x} = \frac{\partial}{\partial x} \left(\frac{Q_s}{w} \right). \quad (5)$$

326 where w is the valley-bottom width. We approximate the mean rate of colluvial deposition at
327 valley heads by

$$328 \quad D\nabla^2 z|_{\text{heads}} \approx \frac{3DS_h}{w}, \quad (6)$$

329 where S_h is the mean slope gradient of toe slopes as they intersect the valley bottom at valley
330 heads (Fig. 5). Equation (6) derives from the mass balance of a square segment of the valley
331 bottom in the vicinity of the valley head (Fig. 5). According to the assumption that soil-mantled
332 hillslopes evolve diffusively, valley bottoms in the vicinity of the valley head receive a unit
333 sediment flux equal to DS_h by colluvial transport processes from each of three adjacent hillslope
334 segments. In the cross-sectional profile, the difference in sediment flux across the profile is equal
335 to DS_h and that difference occurs over a distance of w . As such, the colluvial deposition rate,
336 computed in a manner that is independent of DEM or pixel resolution, includes the valley width
337 in the denominator (Pelletier, 2010a). Similarly, the divergence in the along-valley direction is
338 approximately DS_h/w . The factor of 2 does not appear in the along-valley derivative because
339 colluvial sediment flux enters the valley bottom segment only from the upslope direction.



340 Colluvial sediment flux leaving the valley head is assumed to be negligible because the slope of
341 the valley bottom is typically much smaller than that of the hillslope entering it from upslope.

342 Next, we introduce three power-law relationships that relate volumetric sediment flux to
343 contributing area, channel width to contributing area, and contributing area to distance along
344 flow lines from topographic divides. Sediment flux is a power-law function of contributing area
345 at WGEW (Section 3.1):

$$346 \quad \frac{1}{\varepsilon_0} Q_s = k_{Qs} A^p. \quad (7)$$

347 The coefficient k_{Qs} in equation (7) is a sediment transport efficiency parameter that is a function
348 of runoff, sediment texture, vegetation cover, and potentially other factors. It takes on different
349 values in the shrubland and the grassland (i.e., k_{Qss} and k_{Qsg}), reflecting the fact that sediment
350 yield is a function of vegetation cover. Its value in shrublands, along with that of p , is obtained
351 by a least-squares regression of equation (7) to Q_s values for the five shrubland drainage basins
352 of Lucky Hills reported by Nearing et al. (2007). Its value for the grasslands is constrained by
353 assuming that the same value of p applies to both shrublands and grasslands, and using the single
354 Q_s data point available for grasslands (i.e., for watershed 112) to constrain k_{Qs} using equation (7).

355 Miller (1995) measured 222 cross-sectional channel profiles in the field at WGEW and
356 used those data to calibrate a power-law relationship between channel width and contributing
357 area:

$$358 \quad w = k_w A^l \quad (8)$$

359 where $k_w = 0.023 \text{ m}^{0.32}$ and $l = 0.34$. Contributing area has a power-law relationship with the
360 mean distance, x , from divides along flow lines:

$$361 \quad A = k_A x^c. \quad (9)$$



362 In section 3 we report the values of k_{Qss} and k_{Qsg} , p , k_w , l , k_A , and c for WGEW. Substituting
363 equations (4)-(9) into equation (3) yields

$$364 \quad k_{Qs} k_A^{p-1} x^{c(p-1)} = \frac{k_{Qs}}{k_w} k_A^{p-l} c(p-l) x^{c(p-l)-1} - \frac{3DS_h}{k_w k_A^l x^{cl}}. \quad (10)$$

365 Equation (10) is applied using vegetation-specific values for k_Q and S_h to solve for the mean
366 distance from divides to valley heads in shrublands and grasslands, i.e., x_s and x_g , respectively.

367 That is, x_s is obtained by solving

$$368 \quad k_{Qss} k_A^{p-1} x_s^{c(p-1)} = \frac{k_{Qss}}{k_w} k_A^{p-l} c(p-l) x_s^{c(p-l)-1} - \frac{3DS_{hs}}{k_w k_A^l x_s^{cl}}, \quad (11)$$

369 and x_g is obtained by solving

$$370 \quad k_{Qsg} k_A^{p-1} x_g^{c(p-1)} = \frac{k_{Qsg}}{k_w} k_A^{p-l} c(p-l) x_g^{c(p-l)-1} - \frac{3DS_{hg}}{k_w k_A^l x_g^{cl}}. \quad (12)$$

371 Equation (3) is generally applicable within the fluvial network. Once the colluvial deposition rate
372 is approximated using equation (6) (which makes use of S_h , the mean gradient of the hillslope toe
373 slopes at valley heads), subsequent equations, including equations (10)-(12), apply only to valley
374 heads. Equation (10)-(12) are a mathematical representation of the conceptual model, first
375 proposed by Tarboton et al. (1992), that erosion at valley heads is a competition between
376 transport-limited fluvial erosion and colluvial deposition. That is, fluvial erosion rates must
377 exceed colluvial deposition rates by an amount equal to the net erosion rate on the landscape in
378 order to maintain an equilibrium drainage density.

379

380 **3. Results**

381 **3.1. Topographic analyses**



382 Figure 3 illustrates the results of the drainage network identification for 1.6 km x 1.6 km
383 examples of the landscape in the vicinity of the Lucky Hills and Kendall sites. The mean
384 distance along flow lines from divides to valley bottoms is 18 m in the shrubland area and 50 m
385 in grassland area, using the Pelletier (2013) algorithm. Figures 4A&4B illustrate that hillslopes
386 in the shrubland area are more finely dissected with rills and gullies than in the grassland area.
387 As part of this analysis we also measured the mean slope gradient of hillslopes immediately
388 adjacent to valley heads, i.e. S_h in equation (10). We obtained $S_{hs} = 0.17$ m/m in shrublands and
389 $S_{hs} = 0.19$ m/m in grasslands.

390 Mean hillslope relief increases substantially across the shrubland-to-grassland transition
391 (Fig. 5). Between elevations of approximately 1320 and 1430 m a.s.l., mean relief is uniformly
392 low (approx. 0.5-1 m). Above elevations of approximately 1450 m, hillslope relief increases
393 abruptly and continues to increase with increasing elevation.

394 Contributing area follows a piece-wise power-law function of distance along flow lines
395 from topographic divides (Fig. 7), with one set of values for k_A and c applicable on hillslopes and
396 another set of values applicable within the valley network. Above contributing areas of
397 approximately 50 m^2 (or, equivalently, distances from the divide equal to approximately 15 m),
398 contributing area increases as the 2.5 power of distance from the divide for both grassland and
399 shrubland areas, i.e. $k_A = 0.3 \text{ m}^{-0.5}$ and $c = 2.5$ in equation (9). Below contributing areas of 50 m^2 ,
400 $k_A = 2 \text{ m}^{-0.75}$ and $c = 1.25$. We used $k_A = 0.3 \text{ m}^{-0.5}$ and $c = 2.5$ when solving equations (11) and
401 (12), since these values are most applicable to points within the valley network (i.e. at valley
402 heads and points downstream). This is a self-consistent approach because the solutions to
403 equations (11)&(12) are larger than 15 m (Section 3.2).



404 Plots of mean topographic curvature (i.e. the Laplacian of z) versus contributing area
405 (Fig. 8) indicate that mean curvatures are nearly identical at small and large contributing areas
406 but differ substantially within the range of contributing areas from ~ 30 to 300 m^2 .

407 Plots of mean along-channel slope versus contributing area (Fig. 9) for the shrubland and
408 grassland area follow typical patterns for fluvial topography, i.e. the data follow a power-law
409 relationship for relatively large contributing areas and deviate from power-law scaling at small
410 contributing areas due to the predominance of colluvial processes at hillslope scales. The
411 exponents of the slope-area relationships differ somewhat between the shrubland and grassland
412 areas, i.e. 0.15 for the shrubland and 0.18 for the grassland areas. The slope-area plots also
413 deviate from power-law scaling at different spatial scales. Data from shrublands maintain a
414 power law down to contributing areas of approximately 50 m^2 , while data from the grassland
415 areas deviate at larger spatial scales corresponding to contributing areas of approximately 100
416 m^2 . This finding is consistent with the higher drainage density measured in the shrubland-
417 dominated portions of the landscape compared to the grassland portions.

418 **3.2. Mathematical modeling**

419 In this section we use the results of Section 3.1., together with analyses of the sediment
420 yield reported by Nearing et al. (2007), to constrain the terms in equations (11)&(12) in order to
421 solve for the mean distance from divides to valley heads in shrublands and grasslands. In order to
422 constrain the absolute values of $k_{Q_{ss}}$ and p , we performed a least-squares minimization of
423 equation (11) to the decadal-scale sediment yields reported by Nearing et al. (2007) for
424 watersheds 102-106 (Lucky Hills) (Table 1). This regression yields $k_{Q_{ss}} = 2 \times 10^{-6} \pm \text{m}^{1.56} \text{ yr}^{-1}$ (with
425 a range of values including one standard error from 2×10^{-7} to 2×10^{-5}), $p = 1.44 \pm 0.2$, and $R^2 =$
426 0.93 (Fig. 10). Assuming that the value of p derived from the shrubland watersheds also applies



427 to the grassland watershed, the value of k_{Qs} for the grassland is estimated to be $k_{Qsg} = 6 \times 10^{-8} \text{ m}^{1.56}$
428 yr^{-1} . For D we adopt the value of $1 \times 10^{-3} \text{ m}^2 \text{ yr}^{-1}$ commonly inferred from scarp degradation
429 studies in the southwestern U.S. (e.g., Hanks, 2000, Table 2 cites D values for the Basin and
430 Range of the western U.S. of between 6.4×10^{-4} and $2 \times 10^{-3} \text{ m}^2 \text{ yr}^{-1}$ based on eight published
431 studies). The full list of model parameters and their values is provided in Table 2.

432 We used equations (11)&(12) to predict the mean distance along flow lines from divides
433 to valley bottoms in shrublands and grasslands. The predicted values are $x_s = 16 \text{ m}$ and $x_g = 66 \text{ m}$
434 (Table 3). In the topographic analysis shown in Figure 3, we measured 18 m in shrublands and
435 60 m in grasslands. As such, the model predicts mean distances from divides to valley bottoms
436 within 10% of measured values.

437 Figure 11 plots the magnitude of the three terms for a range of possible mean distances
438 from divide to valley bottom. The fluvial erosion rate is $\sim 10^{-2} \text{ m yr}^{-1}$ in shrublands and $\sim 10^{-3} \text{ m}$
439 yr^{-1} in grasslands, increasing with distance downstream, reflecting the nonlinear relationship
440 between sediment flux and drainage area (Fig. 10). The colluvial deposition rate is $\sim 10^{-3} \text{ m yr}^{-1}$
441 for both shrublands and grasslands and decreases modestly with increasing distance from the
442 divide as a result of the increase in channel width with increasing contributing area (equation
443 (8)). Figure 11 demonstrates that the fluvial erosion rate must be quite large (approximately two
444 orders of magnitude larger than the net erosion rate in this case) in order to counteract the effects
445 of colluvial deposition and thus maintain a valley head.

446 We used the mean curvature versus contributing area data plotted in Figure 8 to
447 reconstruct an average longitudinal profile from divides to valley bottoms in shrublands and
448 grasslands in order to infer the approximate relief reduction associated with the late Holocene
449 shift from grasslands/woodlands to shrublands in WGEW. Figure 12 illustrates the results of this



450 integration. Integrating the curvature versus contributing area data in Figure 8 twice leads to two
451 constants of integration, one of which is constrained by the requirement that the slope along flow
452 pathways at divides is zero and the other by using a constant reference elevation at a contributing
453 area of approximately 300 m^2 . We chose $A \approx 300 \text{ m}^2$ as the location to enforce the reference
454 elevation because this is the contributing area below which the mean curvature begins to deviate
455 significantly between the grassland and shrubland areas. The difference in elevation between the
456 two profiles plotted in Figure 12 provides an estimate of the minimum erosion or relief reduction
457 associated with the shift from grassland to shrubland in the lower elevations of WGEW. We
458 consider the results of Figure 12 to be a minimum estimate because systematic differences in
459 mean slope between the grassland and shrubland across a wide range of scales (Fig. 9) are not
460 reflected (or not fully reflected) in curvature or any reconstruction of the longitudinal profile
461 based on integrating the curvature. The results in Figure 12 suggest that divides have lowered by
462 a minimum of approximately 0.3 m as a result of Holocene vegetation changes. Given hillslope-
463 scale erosion rates of approximately $\sim 10^{-4} \text{ m yr}^{-1}$ in shrublands and the much smaller erosion rate
464 of $\sim 10^{-5} \text{ m yr}^{-1}$ in grasslands, it would take approximately 3 kyr following a transition from
465 grasslands/woodlands to shrublands to erode the landscape by an amount 0.3 m greater in
466 shrublands compared to grasslands/woodlands. This estimate is comparable to the age of the
467 grassland/woodland-to-shrubland transition in the region at an elevation of 1287 inferred from
468 paleovegetation studies in the region, i.e., 2-4 ^{14}C kyr B.P. (Holmgren, 2005).

469

470 **4. Discussion**

471 **4.1. Uncertainty in parameter values and their impact on the model results**



472 Equations (11)&(12) predict mean distances from divides to valley bottoms that are
473 broadly similar to measured values. Several factors limit the accuracy of the comparison between
474 measured and predicted values. First, we relied upon a regional value for the diffusivity D
475 because we do not have a reliable means of calibrating this value locally. Second, decadal-scale
476 erosion rates computed from sediment samplers may differ from long-term erosion rates.
477 Nearing et al. (2007) found that for the six watersheds considered here, half of the sediment yield
478 measured from 1995-2005 was derived from 6-10 events and that the largest events contributed
479 between 9 and 11% of the total yield. Thus, while transport is highly episodic in WGEW, the
480 most effective flood events have a sufficient number of recurrences to provide an estimate of the
481 yield that does not depend sensitively on the time scale. This conclusion is consistent with the
482 fact that erosion rates inferred from sediment sampling from 1995-2005 are similar to erosion
483 rates measured over the post-bomb period using ^{137}Cs (Nearing et al., 2005). That said, the
484 sediment yields reported by Nearing et al. (2007) may not include the most extreme drought
485 conditions or other disturbances that could cause long-term sediment yields to be larger than
486 those reported in Table 1.

487 **4.2. Further discussion of the hypothesis of a vegetation-change-driven increase in drainage** 488 **density in the shrublands of WGEW**

489 We also propose that difference in mean curvatures between shrublands and grasslands
490 between contributing areas of ~ 10 and $\sim 300 \text{ m}^2$ partly reflects recent expansion of the drainage
491 network in the shrublands of WGEW. This hypothesis is consistent with the fact that the
492 deviation of curvature values between the two sites begins at a contributing area comparable to
493 the support area (i.e. the contributing area required to form a valley head) in the grasslands. As
494 such, we propose that the shrubland areas previously had support areas comparable to the



495 grassland areas and that drainage network expansion has influenced the drainage network and the
496 morphology of the adjacent hillslopes down to spatial scales corresponding to contributing areas
497 of $\sim 10 \text{ m}^2$. The fact that curvature values are very similar between shrubland and grassland
498 below spatial scales $\sim 10 \text{ m}^2$ is consistent with the hypothesis that hillslopes in the lower
499 elevations of WGEW have not yet fully adjusted to the increase in drainage density associated
500 with the grassland-to-shrubland transition.

501 Figure 5 demonstrates that mean hillslope relief increases substantially across the
502 shrubland-to-grassland transition. We propose that some of this difference in hillslope relief is a
503 consequence of the difference in fluvial/slope-wash erosion rates, i.e. that higher erosion rates in
504 the lower-elevation shrublands have caused relief reduction in the past few thousand years
505 relative to the higher-elevation grasslands, and that this difference in fluvial/slope-wash erosion
506 rates is the result of a geologically recent increase in drainage density in the shrublands of
507 WGEW. However, it is likely that a portion of the difference in mean hillslope relief across the
508 study site also reflects variable uplift rates, i.e. the fact that the uplift of any piedmont or foothill
509 region tends to increase towards the mountain range. Flexural-isostatic response to erosion
510 (which has been proposed to be an important component of late Cenozoic landscape evolution in
511 southern Arizona (Menges and Pearthree, 1989; Pelletier, 2010b)) of the Dragoon Mountains can
512 be expected to have caused eastward tilting of WGEW, i.e. higher uplift rates in the higher
513 elevations of WGEW compared to the lower elevations. Tilting would not explain the abrupt
514 increase in relief at elevations just above 1430 m a.s.l., however, since no faulting occurs in the
515 vicinity of this contour. Therefore, it is likely that some of the difference in hillslope-scale relief
516 across the shrubland-to-grassland transition at WGEW is a result of the difference in erosion
517 rates between the shrublands and grasslands. While we can be certain of grassland-to-shrubland



518 shift only during the present interglacial period, the Quaternary period has seen many interglacial
519 periods broadly similar in climate to the current period, hence it is likely that the lower
520 elevations of WGEW have seen grassland-to-shrubland conversions more than once over the past
521 approximately 2 Myr. Each of these episodes could have contributed to relief reduction of the
522 lower elevations of the study site relative to the higher elevations.

523 Previous studies at WGEW have emphasized the role of base-level lowering and
524 vegetation changes within the past 130 years on the differences in erosion rate between Lucky
525 Hills and Kendall (Nearing et al., 2007). However, recent paleovegetation studies have provided
526 a new perspective. Specifically, Holmgren's (2005) documentation of shrubland species in the
527 region several thousand years before present at an elevation less than 100 m lower than Lucky
528 Hills suggests that the lower elevations of WGEW likely shifted from a grassland/woodland to a
529 shrubland prior to the 1880s. While base-level lowering has steepened hillslopes and channels
530 close to the main-stem channel of Walnut Gulch, hillslope-scale relief and slope gradients are
531 clearly larger at Kendall than at Lucky Hills (Figs. 6&9), indicating that base-level lowering may
532 be a dominant factor only for those areas within close proximity to the main channel. The
533 magnitude of the differences in topography (i.e., drainage density and the magnitude of erosion
534 than can be inferred from the change) is difficult to fit into a period as short as 130 years. Given
535 erosion rates measured over the past sixty at Lucky Hills, approximately 2 cm of erosion can be
536 expected to have occurred over the past 130 years. Figure 12 suggests that erosion associated
537 with a recent increase in drainage density is likely at least ten times this value, and thus more
538 consistent with a vegetation change that occurred several thousand years before present.

539 **4.3. Implications for our understanding of the controls on drainage density**



540 The model of this paper contributes to our broader understanding of the controls on
541 drainage density and it provides a mathematical model for predicting drainage density that may
542 be useful in other study sites.

543 Previous studies have demonstrated that drainage density is controlled by relief (e.g.
544 Montgomery and Dietrich, 1992; Tarboton, 1992; Tucker and Bras, 1999), climate (Melton,
545 1957; Abrahams and Ponczynski, 1984), parent material (e.g. Ray and Fischer, 1960; Day,
546 1980), and time (e.g. Ruhe, 1952; Dohrenwend et al., 1987). While many studies have
547 demonstrated the importance of individual factors on drainage density, we lack a comprehensive
548 model for drainage density that integrates all of these factors. Equation (10) represents one
549 possible candidate for such a model in soil-mantled, transport-limited landscapes. Time is not
550 included in the model because it is based on an equilibrium mass-balance framework.
551 Nevertheless, equation (10) predicts the drainage density to which a transient landscape will
552 likely approach over time following a perturbation.

553 Relief enters the model via the erosion rate, E (quantified for the case of WGEW using
554 multiple measurements of Q_s/A), and the mean toe slope gradient near valley heads, S_h . Climate
555 and vegetation cover enters the model through the parameters D (which increases with increasing
556 soil moisture and temperature cycling around 0°C (which together drive soil creep) and
557 increasing vegetation cover (which drives bioturbation)) and k_{Q_s} (which increases with rainfall
558 and decreases with vegetation cover)). In addition, equation (10) explicitly includes channel
559 width and its scaling with contributing area, factors that, to our knowledge, have not been
560 included in previous mathematical models of drainage density.

561 Drainage density is most commonly found to be an inverse function of mean annual
562 precipitation or effective precipitation. This finding is consistent with the conceptual model of



563 this paper that vegetation cover is the predominant climate-related variable that influences
564 drainage density, and that vegetation cover and drainage density are inversely related. Melton
565 (1957), for example, documented an inverse correlation between drainage density and the
566 precipitation-effectiveness (P-E) index at over eighty sites in the southwestern United States
567 including arid (low elevation) and humid (high-elevation) climates. A similar negative
568 correlation between drainage density and mean annual precipitation was found by Abrahams and
569 Ponczynski (1984). Naively, one might expect more precipitation to result in greater channel
570 incision and hence less contributing area, between divides and valley heads (and hence a higher
571 drainage density), all else being equal. Melton (1957), however, proposed that greater aridity
572 results in a lower vegetation density and, hence, a reduction in the cohesive strength protecting
573 soils on hillslopes, thus leading to a higher drainage density. One might also expect a lower
574 vegetation density to increase the runoff-to-rainfall ratio and hence also increase drainage
575 density, but runoff intensity varied by only a factor of two across Melton's sites while drainage
576 density varied by nearly two orders of magnitude. Istanbuloglu and Bras (2007) provided
577 theoretical support for Melton's interpretation, illustrating that a lower vegetation densities can
578 lead to higher drainage densities through the cohesive or anchoring effect of plant roots.

579 **4.4. Implications for our understanding of erosion-climate linkages**

580 There has been an ongoing debate in the geomorphic literature regarding the importance
581 of climate (not limited to but often defined as mean annual precipitation (MAP)) on erosion
582 rates. Given the significant correlation between MAP and erosion rates in many studies within
583 individual mountain ranges (e.g., Reiners et al., 2003; Bookhagen and Strecker, 2012), it is
584 perhaps surprising how little correlation exists between MAP and erosion rates in global
585 compilation/synthesis studies (e.g. von Blanckenburg, 2005; Portenga and Bierman, 2011). Even



586 studies that emphasize the climatic control on erosion rates note that R^2 values between erosion
587 rates and MAP are quite small (e.g., Yanites and Kesler, 2015).

588 Recent work on the role of vegetation, and its changes through time, can provide a basis
589 for understanding the relatively low correlation between erosion rates and MAP in global
590 compilation studies and the complex relationship between erosion rates and climate more
591 generally. For example, Torres Acosta et al. (2015) recently documented a negative correlation
592 between erosion rates and both vegetation cover and MAP in Kenya. They proposed that the
593 primary effect of more humid conditions is to increase vegetation cover on hillslopes, thereby
594 reducing erosion rates on otherwise similar slope gradients. This concept is consistent with the
595 classic Langbein and Schumm (1958) curve. Langbein and Schumm (1958) proposed that
596 sediment yields are maximized in semi-arid climates (all else being equal) because such climates
597 generate sufficient rainfall to detach and transport soil in overland/rill flow but insufficient
598 vegetation cover to protect/anchor the soil. As MAP increases in this conceptual model, more
599 precipitation is available to drive erosion, but this effect is more than offset by a decrease in the
600 susceptibility of soil to erode due to the increased anchoring effect associated with greater plant
601 cover/biomass. The results of this paper demonstrate further complexity in the erosion-climate
602 relationship, i.e., that the change in climate (and hence of vegetation cover) can be as important
603 or more important than its mean state. That is, erosion rates can be larger during a humid-to-arid
604 transition than during an arid-to-humid transition, even if the mean climatic conditions (averaged
605 over the transition) are equal. It is important to emphasize that the effect of vegetation on the rate
606 of erosion by colluvial processes may be entirely different than its effect on fluvial/slope-wash
607 processes. All else being equal, increased vegetation cover is likely to increase erosion rates by
608 colluvial processes, since more plants can be expected to drive higher rates of bioturbation (e.g.,



609 Osterkamp et al., 2011). As such, it is crucial to consider colluvial and fluvial/slope-wash
610 processes separately when considering the effects of vegetation on hillslope erosion.

611

612 **5. Conclusions**

613 In this study we leveraged all relevant data from a uniquely well-studied semi-arid
614 watershed to test the hypothesis that late Holocene vegetation changes can modulate drainage
615 density, hillslope-scale relief, and watershed-scale erosion rates. We documented that areas
616 below 1430 m a.s.l. have decadal-scale erosion rates approximately ten times higher, drainage
617 densities approximately three times higher, and hillslope-scale relief approximately three times
618 lower than elevations above 1430 m. We calibrated all the terms of a mathematical landscape
619 evolution model and used the model to predict the equilibrium drainage density associated with
620 shrublands and grasslands. Model predictions for the increase in drainage density associated with
621 the shift from grasslands/woodlands to shrublands are broadly consistent with measured values.
622 Using modern erosion rates and the magnitude of relief reduction associated with the transition
623 from grasslands/woodlands to shrublands, we also estimated the timing of the grassland-to-
624 shrubland transition in the lower elevations of WGEW to be approximately 3 ka, i.e., broadly
625 consistent with constraints from paleovegetation studies. Our work provides a mathematical
626 model for predicting equilibrium drainage density in transport-limited fluvial environments that
627 may be applicable in other study sites.

628

629 **Data Availability**

630 The DEM used in this paper can be obtained upon request from the corresponding author.
631 All other data used in the paper are in the published literature.



632

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636

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- 808 Yanites, B. J., and Kesler, S. E.: A climate signal in exhumation patterns revealed by porphyry
809 copper deposits, *Nature Geos.*, 8, 462–465, doi:10.1038/ngeo2429, 2015.



810 Table 1. Sediment yield data and watershed characteristics. The erosion rate calculation assumes
 811 a soil bulk density of 1500 kg m^{-3}

Watershed ID	Predominant vegetation cover type	Contributing area (ha)	Sediment yield ($\text{t ha}^{-1} \text{ yr}^{-1}$)	Erosion rate (mm yr^{-1})
102	Shrub	1.46	2.31	0.154
103	Shrub	3.68	5.66	0.377
104	Shrub	4.53	1.36	0.091
105	Shrub	0.18	0.75	0.050
106	Shrub	0.34	0.80	0.053
112	Grass	1.86	0.07	0.005

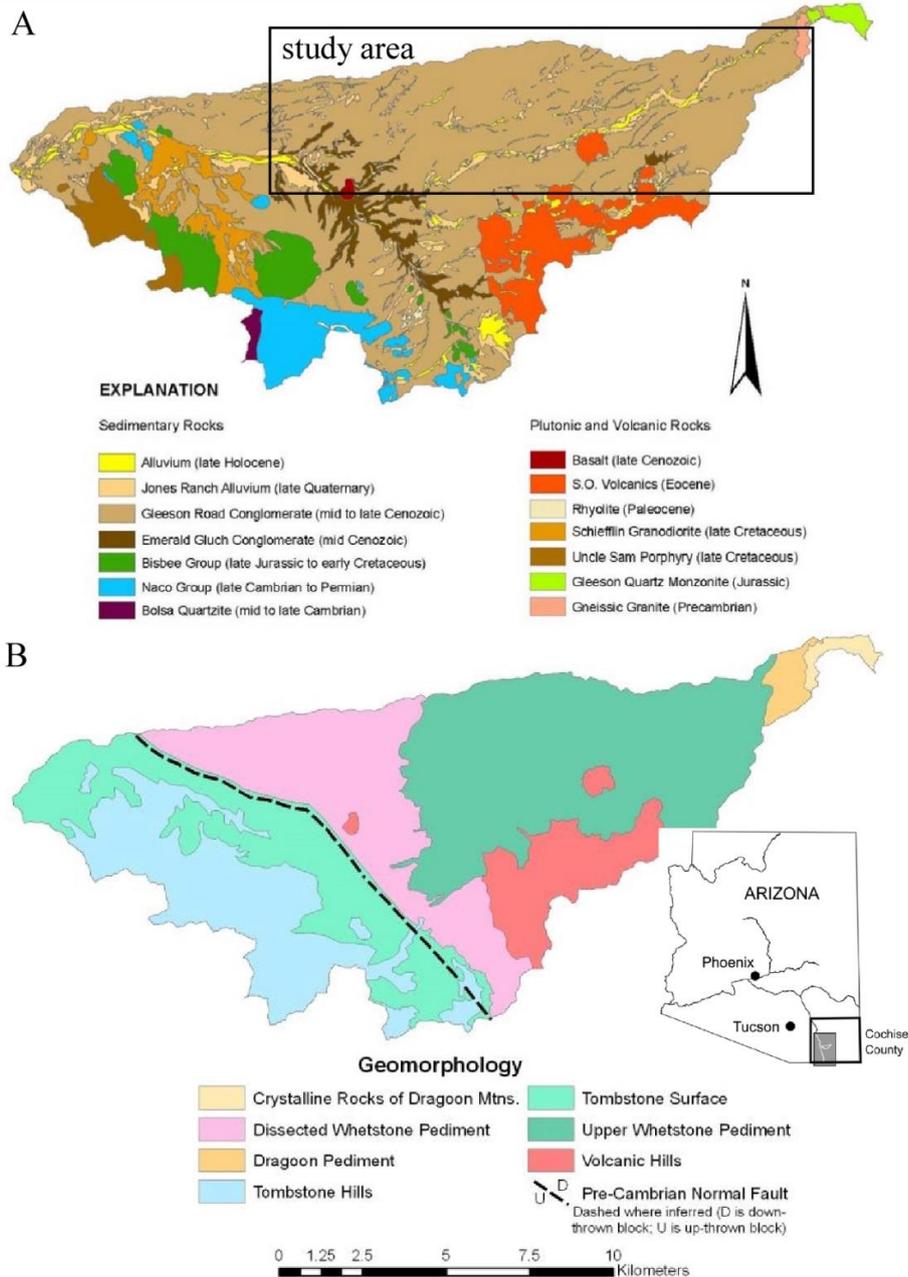
812
 813 Table 2. List of model parameters and values.

Symbol	Units	Description	Value
D	$\text{m}^2 \text{ yr}^{-1}$	topographic diffusivity	1×10^{-3}
S_{hs}	unitless	mean gradient of toe slopes (shrublands)	0.17
S_{hg}	unitless	mean gradient of toe slopes (grasslands)	0.19
l	unitless	exponent of width-area relationship	0.34
k_w	$\text{m}^{0.32}$	coefficient of width-area relationship	0.023
c	unitless	exponent of area-distance relationship	2.5
k_A	$\text{m}^{-0.5}$	coefficient of area-distance relationship	0.3
p	unitless	exponent of sediment-flux-area relationship	1.44
k_{Qss}	$\text{m}^{1.56} \text{ yr}^{-1}$	sediment transport coefficient (shrublands)	2×10^{-6}
k_{Qsg}	$\text{m}^{1.56} \text{ yr}^{-1}$	sediment transport coefficient (grasslands)	6×10^{-8}

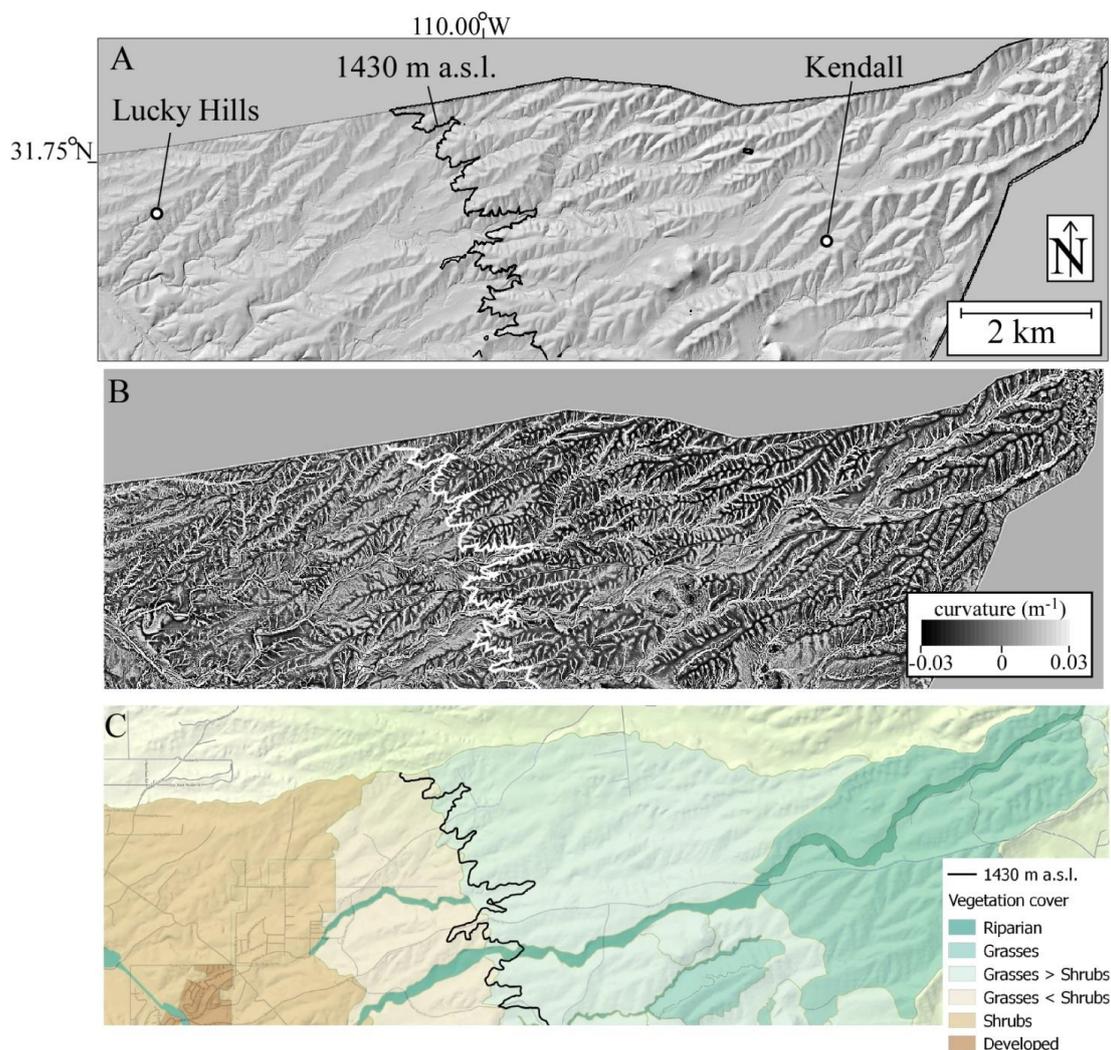
814
 815 Table 3. Measured (from DEM analysis) and predicted values (from equations (11)&(12)) for the
 816 mean distance from divides to valley bottoms in shrublands (x_s) and grasslands (x_g).

	Measured (m)	Predicted (m)
x_s	18	16
x_g	60	66

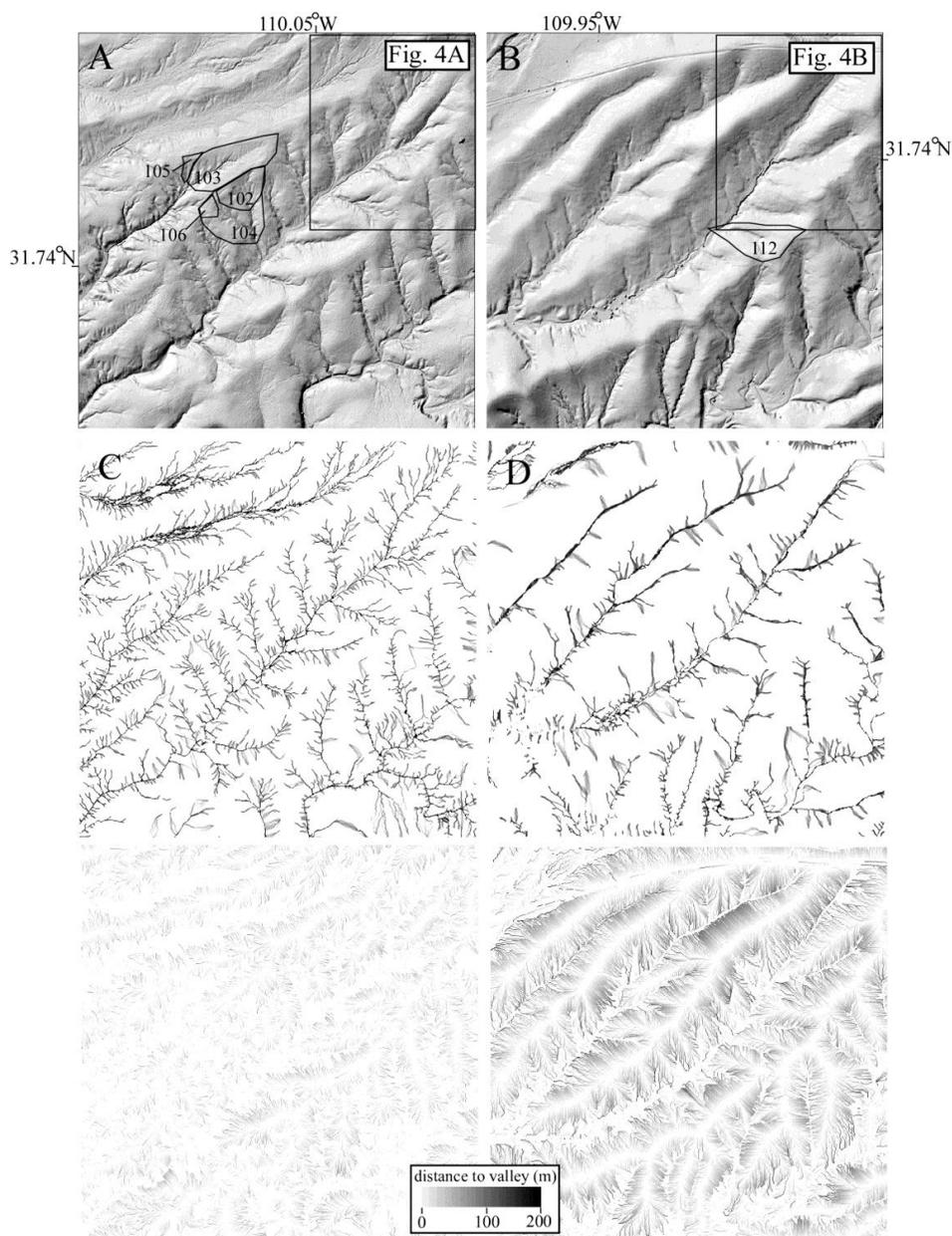
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818
 819 Figure 1. Maps of the bedrock geology and geomorphology of Walnut Gulch Experimental
 820 Watershed (WGEW). (A) Bedrock geology from Osterkamp (2008). Rectangle identifies the
 821 portion of WGEW that is the focus of this study. (B) Geomorphic map from Osterkamp (2008).
 822 The boundary between the Dissected Whetstone Pediment and Upper Whetstone Pediment
 823 marks a key transition in landscape morphology, soil type, and vegetation cover (Fig. 2).
 824



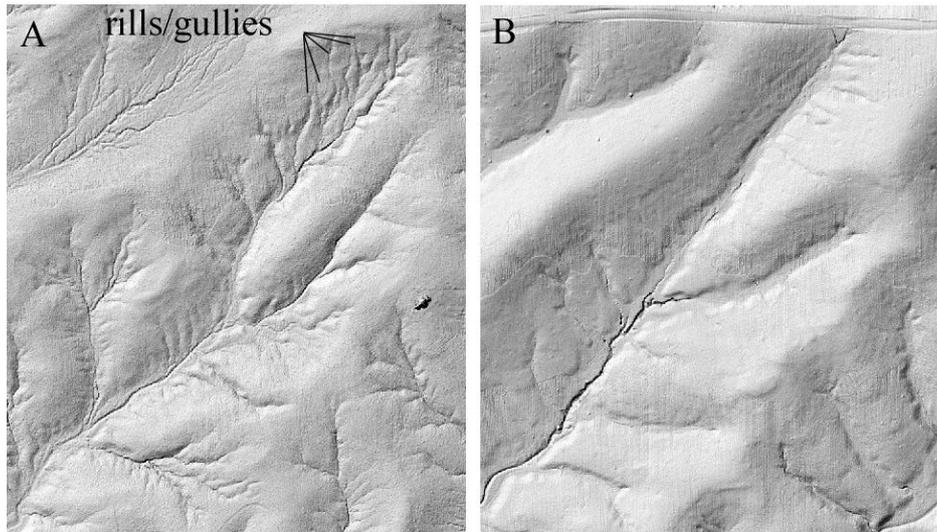
825
826 Figure 2. Relationships among landscape morphology and vegetation cover in the study area. (A)
827 Shaded relief image of the topography, illustrating the significant increase in hillslope-scale
828 relief from the western to the eastern portion of the study area. (B) Grayscale map of topographic
829 curvature (i.e., Laplacian), demonstrating generally lower absolute hillslope curvatures (i.e.,
830 more gray) in the western (shrubland) portion of the study area relative to the eastern (grassland)
831 portion (more black). (C) Vegetation map, after Skirvin et al. (2008), identifying the areas that
832 are primarily shrubland, grassland, and transitional between the two.
833



834
835 Figure 3. Drainage density is higher in shrubland areas than in grassland areas of the study site.
836 Shaded relief images of representative 1.6 km x 1.6 km areas of (A) shrublands, including Lucky
837 Hills watersheds 102-106 and (B) grasslands, including Kendall watershed 112. (C)&(D) Images
838 of the drainage network identified using the Pelletier (2013) algorithm for the areas shown in
839 (A)&(B), respectively. (E)&(F) Grayscale maps of the distance along flow lines from divides to
840 the valley bottom. The mean of the mapped values in (E) at valley heads is 18 m while the mean
841 of the mapped value in (F) at valley heads is 60 m.



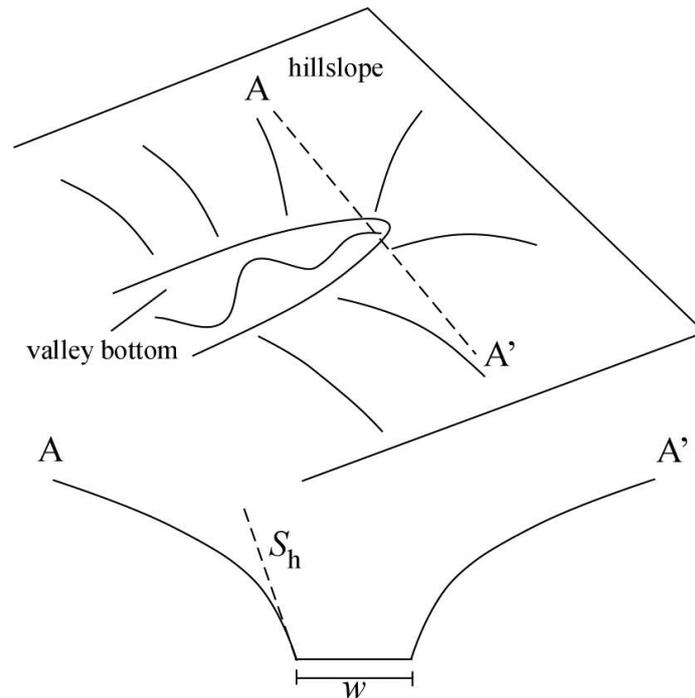
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843

844 Figure 4. Detailed shaded-relief images (locations shown in Fig. 3) illustrating the presence of
845 parallel hillslope rills and gullies in the shrubland areas (shown in A). Grassland areas (shown in
846 B) have fewer such features.

847

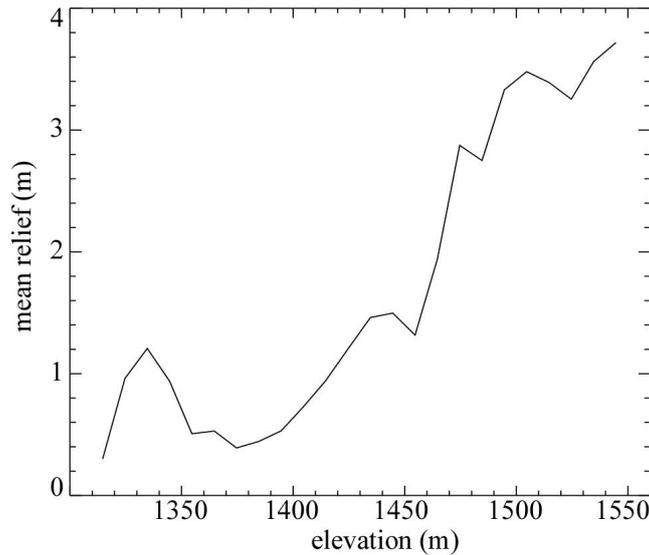


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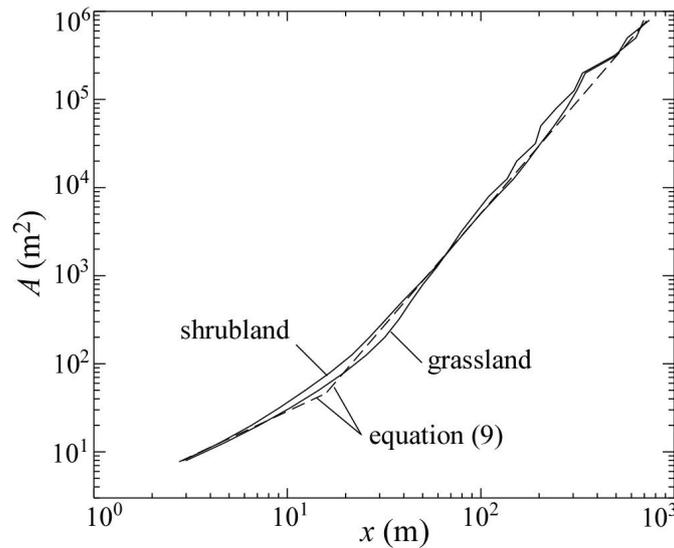
849 Figure 5. Schematic figure of a valley head. The profile shown along A-A' is the cross-section of
850 the valley head where colluvial sediment flux from hillslopes of mean gradient S_h enter a
851 segment of width w . The requirement that fluvial erosion rates must be greater than colluvial



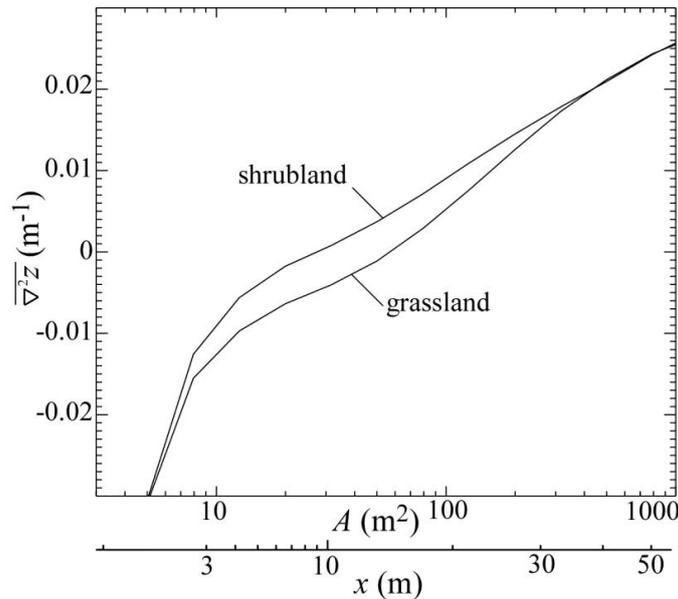
852 deposition rates at the valley head provides a quantitative criterion for predicting the drainage
853 density of landscapes.
854



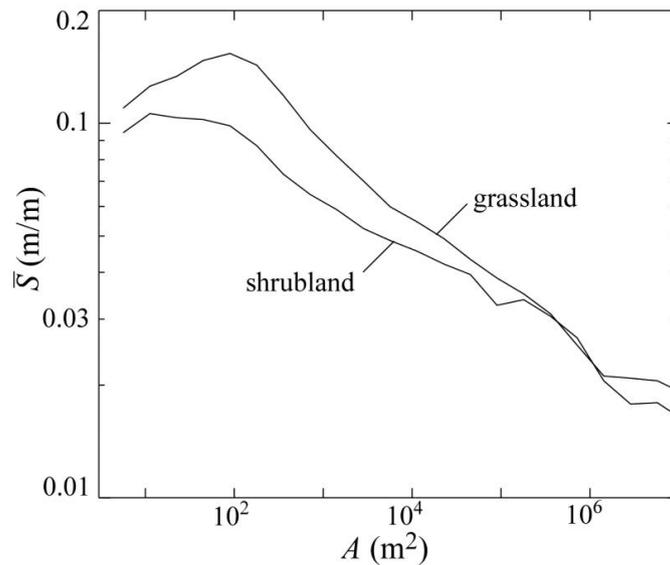
855
856 Figure 6. Plot of mean hillslope relief as a function of elevation, illustrating the marked increase
857 in relief above elevations of approximately 1430 m a.s.l. in the study area. Each data point
858 represents the mean hillslope relief in 10-m-wide elevation bins.
859



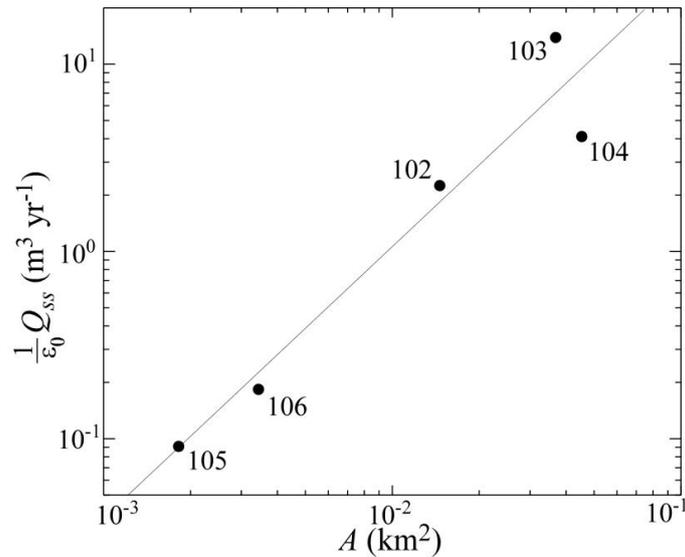
860
861 Figure 7. Plots of contributing area versus mean distance from the divide for the shrubland
862 (approximated as the portion of the study area below 1430 m a.s.l.) and grassland areas (above
863 1430 m). The dashed line plots the piece-wise power-law relationship (equation (9)) exhibited by
864 the data.
865



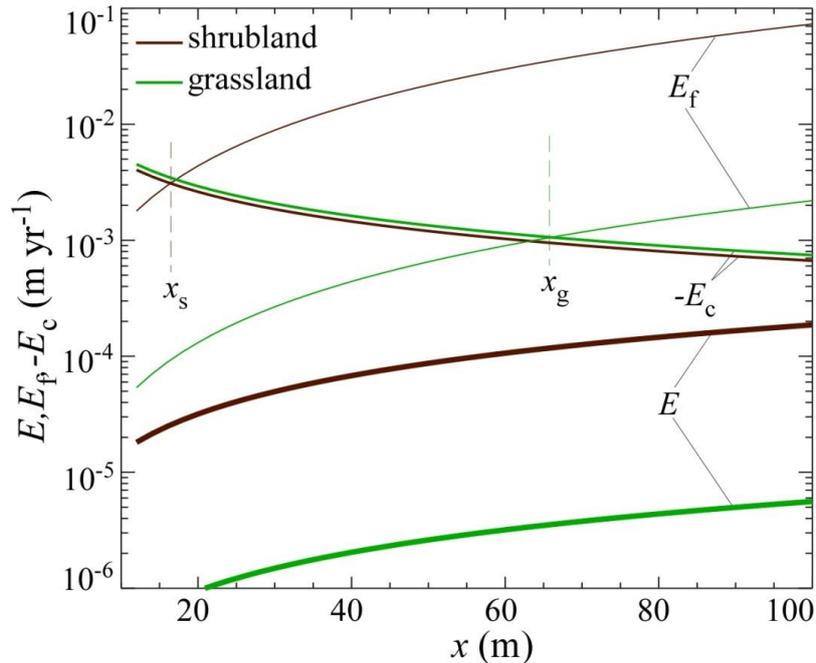
866
 867 Figure 8. Plots of mean topographic curvature as a function of contributing area (distance from
 868 divide also shown along x axis using data in Fig. 7). Topographic curvatures are similar in
 869 shrublands and grasslands at small and large contributing areas, with a minimum of 0.03 m^{-1} near
 870 divides. Within a range of contributing areas from ~ 10 to $\sim 300 \text{ m}^2$ the data show significant
 871 differences in mean curvature between shrublands and grasslands.
 872



873
 874 Figure 9. Plot of mean slope versus contributing area. Shrublands and grasslands both show a
 875 power-law relationship, with deviations from power-law behavior occurring at larger spatial
 876 scales in grasslands relative to shrublands.
 877



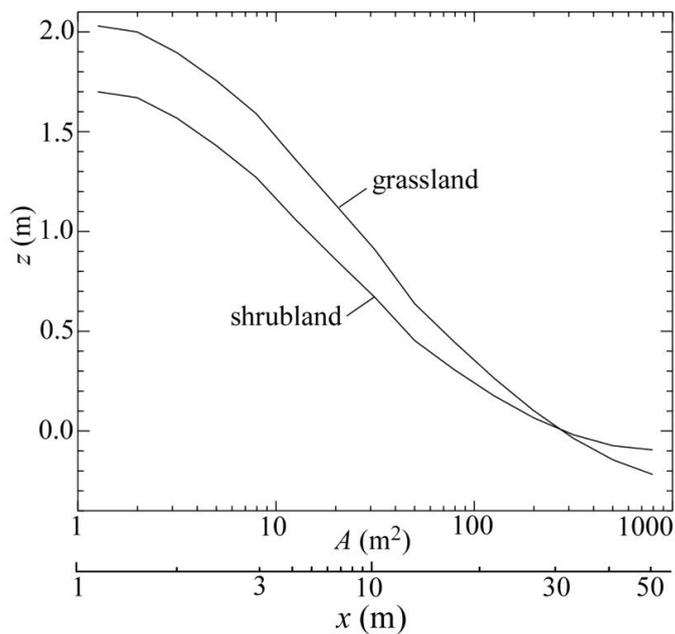
878
 879 Figure 10. Plot of the decadal time-scale volumetric sediment fluxes in shrublands measured at
 880 the five watersheds of the Lucky Hills as a function of contributing area. The straight line is the
 881 result of a least-squares regression to the logarithms of both sides of equation (7), from which the
 882 values of $k_{Q_{ss}}$ and p were constrained.
 883



884
 885 Figure 11. Plots of the total erosion rate, E , the fluvial/slope-wash erosion rate, E_f , and the
 886 colluvial deposition rate, $-E_c$, as a function of distance along flow lines from divides, x . The
 887 values of x_s and x_g (where $E = E_f + E_c$) are also shown.



888



889

890 Figure 12. Plots of the mean longitudinal profile of hillslopes and valley bottoms in shrubland
891 and grassland areas, constructed by integrating the mean curvature data in Figure 8.