



Effects of seasonal variations in vegetation and precipitation on catchment erosion rates along a climate and ecological gradient: Insights from numerical modelling

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7 Abstract. Precipitation in wet seasons influences catchment erosion and contributes to annual erosion rates. However, wet 8 seasons are also associated with increased vegetation cover, which helps resist erosion. This study investigates the effect of 9 present-day seasonal variations in rainfall and vegetation cover on erosion rates for four catchments along the extreme climate 10 and ecological gradient (from arid to temperate) of the Chilean Coastal Cordillera ($\sim 26 \text{ }^\circ\text{S} - \sim 38 \text{ }^\circ\text{S}$). We do this using the 11 Landlab-SPACE landscape evolution model modified to account for vegetation-dependent hillslope-fluvial processes and 12 hillslope hydrology. Model inputs include present-day (90 m) topography, and a timeseries (from 2000-2019) of MODISderived NDVI for vegetation seasonality; weather station observations of precipitation; and evapotranspiration obtained from 13 14 GLDAS NOAH. Simulations were conducted with a step-wise increase in complexity to quantify the sensitivity of catchment 15 scale erosion rates to seasonal variations in precipitation and/or vegetation cover. Simulations were conducted for 1,000 years (20 years of vegetation and precipitation observations repeated 50 times). After detrending the results for long-term transient 16 17 changes, the last 20 years were analyzed. Results indicate that when vegetation cover is varied but precipitation is held constant, 18 the amplitude of change in erosion rates relative to mean erosion rates ranges between 6.5% (humid-temperate) to 36% 19 (Mediterranean setting). In contrast, in simulations with variable precipitation change and constant vegetation cover, the 20 amplitude of change in erosion rates is higher and ranges between 13% (arid) to 91% (Mediterranean setting). Finally, 21 simulations with coupled precipitation and vegetation cover variations demonstrate variations in catchment erosion of 13% 22 (arid) to 97% (Mediterranean setting). Taken together, we find that precipitation variations more strongly influence seasonal 23 variations in erosion rates. However, the effects of seasonal variations in vegetation cover on erosion are also significant 24 (between 5-36%) and are most pronounced in semi-arid to Mediterranean settings and least prevalent in arid and humid-25 temperature settings.

26 Keywords: Landlab, vegetation, Chilean Coastal Cordillera, biogeomorphology, seasonality, precipitation.

27 1 Introduction

28 Catchment erosion rates vary spatially and temporally (e.g., Wang et al., 2021) and depend on topography (slope, Carretier et 29 al., 2018) vegetation cover and type (e.g., Zhang et al., 2011; Starke et al., 2020; Schaller and Ehlers, 2022) and precipitation 30 rates (e.g., Cerdà, 1998; Tucker and Bras, 2000). Over annual timescales, temporal variations in catchment erosion occur in 31 response to seasonal variations in precipitation and vegetation cover. For example, previous work has found that a significant 32 fraction of annual erosion occurs during wet seasons, with high runoff rates (Hancock and Lowry, 2021; Leyland et al., 2016; 33 Gao et al., 2021; Wulf et al., 2010). However, this increase in precipitation during wet seasons also promotes vegetation 34 growth, which in turn influences erosion rates (Langbein and Schumm, 1958; Zheng, 2006; Schmid et al., 2018). Seasonality, 35 and longer-term, changes in both precipitation and vegetation cover plays a crucial role in intra-annual changes in erosion rates 36 (Istanbulluoglu and Bras, 2006; Yetemen et al., 2015; Schmid et al., 2018; Sharma et al., 2021). The intensity, frequency, and 37 seasonality of precipitation and vegetation cover change within a year depend on the climate and ecological conditions of the

area of interest (Herrmann and Mohr, 2011). One means of investigating the effects of seasonality in precipitation and (or)





39 vegetation cover on erosion rates is through landscape evolution modeling (LEM), which can be parameterized for variations 40 in vegetation-dependent hillslope and fluvial processes over seasonal time scales. 41 Previous modeling and observational studies have investigated the effects of seasonality in precipitation and vegetation on 42 catchment erosion. Bookhagen et al., (2005), Wulf et al., (2010), and Deal et al., (2017) investigated the effects of stochastic 43 variations in precipitation on erosion and sediment transport in the Himalayas. They found that high variability in rainstorm 44 days (>80% of MAP) during the wet season (summer monsoon) caused high variability in the suspended sediment load. Work 45 by Chakrapani (2005) identifies the control of mean local relief and seasonality in precipitation on sediment load in rivers. 46 Similar seasonality in sediment loads was reported in a field study in Iran, using sediment traps and erosion pins. These authors 47 concluded that wet seasons have maximum erosion rates (>70% of annual), which decreases in the dry season (<10% of annual) 48 (Mosaffaie et al., 2015). Field observations in the heavily vegetated Columbian Andes concluded that soil erosion and nutrient 49 losses are significantly influenced by precipitation seasonality (Suescún et al., 2017). In contrast, work by Steegen et al., (2000) 50 in a loamy agricultural catchment in central Belgium found suspended sediment concentrations in streams were lower during 51 summer (wet) rather than winter (dry) months due to the development in vegetation cover in the wet season. Other workers 52 have found a dependence of seasonal erosion on ecosystem type. For example, Istanbulluoglu et al., (2006) found a reduction 53 in soil loss potential to storm frequency in humid ecosystems compared to arid and semi-arid regions. Work by Wei et al., 54 (2015) documented that differences in vegetation cover may contribute to long-term erosion and sedimentation. However, 55 seasonal variations in runoff and sediment yield are mainly influenced by intra-annual rainfall variations. Finally, previous 56 work in a Mediterranean environment (Gabarrón-Galeote et al., (2013) described rainfall intensity as the main factor in 57 determining hydrological erosive response, regardless of the rainfall depth of an event. 58 When looking at seasonal vegetation changes in more detail, several different studies suggest these changes are important for 59 catchment erosion. For example, Garatuza-Payán et al., (2005) emphasized that seasonal patterns in erosion are strongly 60 influenced by plant phenology demonstrated by the changes in vegetation cover (as measured by NDVI). A similar study on 61 the Loess Plateau, China, by Zheng (2006) documented decreasing soil erosion as vegetation cover increases during the wet 62 season. Work conducted in a forested setting (Zhang et al., 2014) documented the importance of tree cover as an effective 63 filter for decreasing the effects of rainfall intensity on soil structure, runoff, and sediment yield. Numerical modeling studies 64 have also found a significant impact of vegetation on erosion. For example, Zhang et al., (2019) found that when precipitation 65 is kept constant, the effect of vegetation cover change on sediment yields is significant (20-30% of the total flux). Also, during 66 the early to the mid-wet season, the richness and evenness of plant cover play an essential role in reducing erosion rates during 67 low rainfall events (Hou et al., 2020). However, in the case of high-intensity rainfall events at the start of a wet season, when 68 vegetation cover is low, the duration and intensity of rainfall were found to significantly affect erosion rates (Hancock and 69 Lowry, 2015). Other work conducted in a Mediterranean environment points to the coincidence of peak rainfall erosivity in 70 low vegetation cover settings, leading to an increased risk of soil erosion (Ferreira and Panagopoulos, 2014). Despite 71 potentially conflicting results in the previous studies, what is clear is that seasonality in precipitation and vegetation covers are 72 co-conspire to influence catchment erosion, although which factor (precipitation or vegetation) plays the dominant role is 73 unclear. 74 This study complements the previous work by applying a Landscape Evolution Model (LEM) to investigate seasonal transients 75 in catchment erosion due to variations in precipitation and vegetation. We do this for four locations in the extreme climate and

ecological gradient (i.e., arid, semi-arid, Mediterranean, and humid temperate) of the Chilean Coastal Cordillera. Our efforts are focused on testing two hypotheses: (1) if precipitation is the first-order driver of seasonal erosion rates, then the influence of seasonal changes in vegetation cover would be of low significance, and (2) catchment erosion in arid and semi-arid regions is more sensitive to seasonality in precipitation and vegetation than the Mediterranean and humid temperate regions. To test the above hypotheses, we conduct a sensitivity analysis of fluvial and hillslope erosion over four Chilean study areas to

81 investigate the individual effects of seasonal changes in vegetation cover and precipitation compared to simulations with

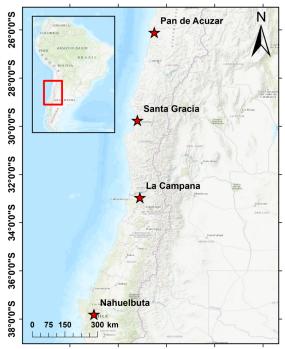




coupled variations in precipitation and vegetation cover. We do this using a two-dimensional LEM (the Landlab-SPACE software), which explicitly handles bedrock and sediment entrainment and deposition. We build upon the approach of Sharma et al., (2021) with the additional consideration of soil-water infiltration. Our model setup is focused on present conditions in the Chilean Coastal Cordillera (Fig. 1) and use as input present day topography from SRTM DEMs (90 m) for four regions with different climate/ecological settings. Simulations in these different ecosystems are drive by observed variations vegetation cover from MODIS NDVI (between 2000 – 2019) and observed precipitation rates over the same time period from neighboring weather stations.

89 2 Study Areas

- 90 This section summarizes the geologic, climate, and vegetation settings of the four selected catchments (Fig. 1) investigated in
- 91 the Chilean Coastal Cordillera. These catchments (from north to south) are located in the Pan de Azúcar National Park (arid,
- 92 ~26°S), Santa Gracia nature preserve (semi-arid, ~30°S), and the La Campana (mediterranean, ~33°S) and Nahuelbuta
- 93 (temperate-humid, \sim 38°S) national parks. Together, these study areas span \sim 1,300 km distance of the Coastal Cordillera. These
- study areas are chosen for their steep climate and ecological gradient from North (arid environment with small shrubs) to South
- 95 (humid temperate environment with evergreen mixed forests) (Schaller et al., 2020). The study areas are part of the German-
- 96 Chilean priority research program EarthShape (<u>www.earthshape.net</u>) and ongoing research efforts within these catchments.



97

74°0'0"W 72°0'0"W 70°0'0"W 68°0'0"W 66°0'0"W

Figure 1. Study areas in the Coastal Chilean Cordillera ranging from an arid environment in the North (Pan de Azúcar), semi-arid (Santa Gracia), Mediterranean (La Campana), and humid temperate environment in the South

Azúcar), semi-arid (Santa Gracia), Mediterranean (La Campana), and humid temperate environment in the South
 (Nahuelbuta). The above map is obtained from the Environmental System Research Institute (ESRI) map server

- 101 (https://services.arcgisonline.com/ArcGIS/rest/services/World Topo Map/MapServer, last access: 25 April 2022).
- 102 The bedrock of the four study areas is composed of granitoid rocks, including granites, granodiorites, and tonalites in Pan de
- 103 Azúcar, La Campana, and Nahuelbuta, respectively and gabbro and diorites in Santa Gracia (Oeser et al., 2018). The soil types





104 in each catchment were identified as a sandy loam in three northern catchments (with high bulk density: 1300 - 1500 kg m³) and sandy clay loam in Nahuelbuta (with lower bulk density: 800 kg m⁻³) (Bernhard et al., 2018). The western margin of Chile 105 106 along the latitudes of the different study areas is characterized by a similar tectonic setting whereby an oceanic plate (currently the Nazca plate) has been subducting under the South American plate since the Palaeozoic. Despite this common tectonic 107 108 setting along, slight differences in modern rock uplift rates are documented in the regions surrounding the three northern 109 catchments (i.e., < 0.1 mm yr⁻¹ for ~ 26 °S to ~ 33 °S) (Melnick, 2016) and the southern catchment (i.e., 0.04 to > 0.2 mm yr⁻¹ 110 for ~38 °S over the last 4±1.2 Ma) (Glodny et al., 2008; Melnick et al., 2009). Over geologic (millennial) timescales, measured denudation rates in the region range between ~0.005 to ~0.6 mm yr-1 (Schaller et al., 2018). To facilitate a comparison between 111 112 the study areas and focus on erosion variations from seasonal changes in precipitation and vegetation, we assume a uniform rock uplift rate of 0.05 mm yr⁻¹ for this study. This rate is broadly consistent with the range of previously reported values. 113

114 The climate gradient in the study areas ranges from an arid climate in Pan de Azúcar (north) with mean annual precipitation

115 (MAP) of \sim 11 mm yr⁻¹ to semi-arid in Santa Gracia (MAP: \sim 88 mm yr⁻¹), the Mediterranean in La Campana (MAP: \sim 350 mm

 $116 yr^{-1}$), and a temperate-humid climate in Nahuelbuta (south) with a MAP of 1400 mm yr^{-1} (Ziese et al., 2020). The observed

117 mean annual temperatures (MAT) also vary with latitude ranging from $\sim 20^{\circ}$ C in the north to $\sim 5^{\circ}$ C in the south (Übernickel et

al., 2020). The previous gradients in MAP and MAT and latitudinal variations in solar radiation result in a southward increase

119 in vegetation density (Bernhard et al., 2018). The vegetation gradient evident from mean MODIS Normalized Difference

120 Vegetation Index (NDVI) values range from ~0.1 in Pan de Azúcar (north) to ~0.8 in Nahuelbuta (south) (Didan, Kamel,

2015). In this study, NDVI values are used as a proxy for vegetation cover density, similar to the approach of Schmid et al.(2018).

123 This gradient in climate and vegetation cover from north to south in the Chilean Coastal Cordillera provides an opportunity to

124 study the effects of seasonal variations in vegetation cover and precipitation on catchment-scale erosion rates in different

125 environments.

126 3 Methods

127 This section comprises a description of model inputs (section 3.1), estimation of runoff rates (section 3.2), model setup (section

128 3.3), and initial and boundary conditions (section 3.4). This is followed by an overview of simulations conducted (section 3.5)

129 and a brief description of how detrending the model results was conducted to remove long-term transients (section 3.6).

130 **3.1 Data used for model inputs**

In contrast to previous modeling studies (Schmid et al., 2018; Sharma et al., 2021) in the same regions, we used present-day 131 132 topography as the initial condition for simulations instead of a synthetic topography produced during a model spin-up phase 133 in LandLab. Initial topography for the four selected catchments was obtained by cropping the SRTM digital elevation model (DEM) in rectangular shapes encapsulating the catchment of interest (Fig. 1). These catchments are same as those investigated 134 135 with previous soil, denudation, and geophysical studies (e.g., Bernhard et al., 2018; Oeser et al., 2018; Schaller et al., 2018; 136 Dal Bo et al., 2019)/ The DEM has a spatial resolution of 90 m and is the same as the cell size used in the model (dx and dy) (SRTM data set of Earth Resources Observation And Science (EROS) Center, 2017). Maximum relief of ~1852 m is observed 137 in La Campana (~33 °S), followed by ~1063 m in Santa Gracia (~30 °S), ~809 m in Nahuelbuta (~38 °S) and ~623 m Pan de 138 Azúcar (~26 °S). Catchment sizes considered vary between ~64 km² in Pan de Azúcar, ~142.5 km² in Santa Gracia, ~106.8 139 140 km² in La Campana and ~68.7 km² in Nahuelbuta. We note that present-day topography as the initial condition in simulations can introduce an initial transient in erosion rates due to model erosional parameters (e.g., erodibility, hillslope diffusivity) 141 142 differing from actual parameters within the catchment, we address this issue through a detrending of model results described 143 later.





144 Precipitation data applied over each study area (Fig. 3b) was acquired from the Global Precipitation Climatology Centre (GPCC) for the period 01/03/2000 to 31/12/2019 (DD/MM/YEAR). The data has a spatial resolution of 1° (~111 km) and a 145 146 1-day temporal resolution and comprises daily land-surface precipitation from rain gauges built on Global Telecommunication System-based and historic data (Ziese et al., 2020). The previous data was augmented with daily precipitation weather station 147 148 data from 01/02/2020 to 28/02/2020 obtained from Übernickel et al., (2020). We do this to include all the seasons between 149 2000 to 2019, i.e., from the Austral Autumn of 2000 to the Austral Summer of 2019. The periods (months of a year) of specific 150 seasons in the Chilean Coastal Cordillera are illustrated in Table 1. Seasonal precipitation rates were calculated by summing up daily precipitation rates at three-month intervals. The seasonality and intensity of precipitation in the wet season (winter) 151 152 increases from the arid (Pan de Azúcar) to humid temperate (Nahuelbuta) region.

153 Table 1. Months of a year corresponding to specific seasons in the Chilean Coastal Cordillera

Seasons	Months
Summer ^{d*}	December - February
Autumn ^{w*}	March - May
Winter ^{w*}	June - August
Spring ^{d*}	September - November

154 155

*d: dry season, w: wet season

NDVI derived from remote sensing imagery has been proven as an effective tool to estimate seasonal changes in vegetation

157 cover density (Garatuza-Payán et al., 2005). Normalized difference vegetation index (NDVI) values were obtained from

MODIS (Didan, Kamel, 2015) satellite data and were used as a proxy for changes in vegetation cover in the catchments. The NDVI data were acquired for 20 years (01/03/2000 - 28/02/2020), with a spatial resolution of 250 m and temporal resolution

NDVI data were acquired for 20 years (01/03/2000 - 28/02/2020), with a spatial resolution of 250 m and temporal resolution of 16 days. For application within the model simulations, the vegetation cover dataset was resampled to match the spatial

resolution (90 m) of SRTM DEM and temporal resolution of 3 months. To summarize, season variations in precipitation rate

and vegetation cover were applied to the simulations between 01/03/2000 and 28/02/2020 and encompass a 20-year record of

163 observation variations in these factors.

164 Additional aspects of the catchment hydrologic cycle were determined using the following approaches for the same time period

165 previously mentioned. First, evapotranspiration (ET) data was obtained from Global Land Data Assimilation System (GLDAS)

166 Noah version 2.1, with a monthly temporal resolution and spatial resolution of 0.25° (~28 km) (Beaudoing et al., 2020; Rodell

et al., 2004). The data was obtained from March-2000 to February-2020. Due to the coarse resolution of the dataset, ET is
 assumed to be uniform over the entire catchment area. No higher resolution datasets were available over the 20-year time-

169 period of interest.

170 Soil properties such as the grain size distribution (sand, silt, and clay fraction) and bulk density were adapted from Bernhard

et al., (2018) to estimate soil water infiltration capacity in each study area. Based on these soil properties, the soils have been

172 classified as a sandy loam (in Pan de Azúcar, Santa Gracia, and La Campana) and sandy clay loam (Nahuelbuta). Average

bulk density values of 1300 kg m⁻³, 1500 kg m⁻³, 1300 kg m⁻³, and 800 kg m⁻³ were used for Pan de Azúcar, Santa Gracia, La

174 Campana, and Nahuelbuta, respectively (Bernhard et al., (2018).



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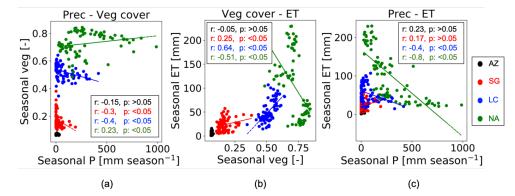


Figure 2. Parameter correlation for model input data (i.e., seasonal precipitation, vegetation cover and evapotranspiration) including: (a) seasonal precipitation [mm season⁻¹] and fraction of vegetation cover [-], (b) fractional seasonal vegetation cover [-] and evapotranspiration [mm], and (c) seasonal precipitation [mm season⁻¹] and evapotranspiration [mm]. The plots represent observations corresponding to Autumn of 2000 to Summer of 2019. Each data point represents one season and are color coded by study area (AZ: Pan de Azúcar, SG: Santa Gracia, LC: La Campana, and NA: Nahuelbuta).

182 Figure 2 shows correlations between the model input data, such as variable climatic or hydrologic cycle metrics (i.e., 183 precipitation and evapotranspiration) and vegetation cover for each study area investigated. The relationships, and regression 184 lines, shown for each area in different climate-ecological zones the general seasonal relationships over the 20 years (i.e., 185 Autumn of 2000 - Summer of 2019) of data. For example, the correlation between seasonal precipitation and vegetation cover 186 (Fig. 2a) illustrates a moderate negative correlation (r > -0.4) in the semi-arid (SG) and Mediterranean (LC) regions. In contrast, 187 vegetation in the humid temperate region (NA) is positively correlated (r: 0.23). ET and vegetation cover are positively 188 correlated (Fig. 2b) in the semi-arid (SG) and Mediterranean regions (LC). However, the correlations are negative in the humid-189 temperate region (NA). The correlation between seasonal precipitation and ET (Fig. 2c) is slightly positive (r: ~ 0.2) in the 190 semi-arid region (SG) and moderately negative (r: -0.4) in the Mediterranean (LC) study area. However, we observe a strong 191 negative correlation (r: -0.8) between precipitation and ET in humid-temperate and Mediterranean regions (LC and NA, Fig. 192 2c). This negative correlation is owed to the steep negative gradient in temperature (e.g., ~ 2.5 °C in NA) and solar radiation 193 (Übernickel et al., 2020) during winters (wet season) in southern latitudes (LC and NA).

194 **3.2 Estimation of runoff rates**

The precipitation rates $[m \text{ season}^{-1}]$ are subjected to soil-water infiltration $[m \text{ season}^{-1}]$ and evapotranspiration $[m \text{ season}^{-1}]$ to estimate the seasonal runoff rates $[mm \text{ season}^{-1}]$. The runoff rates (R) at every time step (t) are calculated using actual soilwater infiltration (I_a) and evapotranspiration (ET) as follows,

198
$$R(t) = P(t) - I_a(t) - ET(t)$$
(1)

where, *P* is the precipitation amount in a season. This relationship was applied in the model grid cells with non-zero sedimentthickness.

The soil-water infiltration rate was estimated by applying the Green-Ampt equation (Green and Ampt, 1911; Julien et al., 1995):

203
$$f(t) = K_e \left(1 + \frac{\psi \cdot \Delta \theta}{F} \right)$$
(2)





- where f(t) is the infiltration rate [m s⁻¹] at time t, K_e is the effective hydraulic conductivity [m s⁻¹], F is the cumulative infiltration [m], Ψ is the suction at the wetting front [m], and $\Delta\theta$ is the difference between saturated and initial volumetric moisture content [m³ m⁻³]. Effective hydraulic conductivity is highly variable and anisotropic; hence, it was considered to be uniform with a value of 1×10^{-6} m s⁻¹ for each catchment. Following the approach of Istanbulluoglu and Bras, (2006) for loamy soils, the soil-water infiltration was modified to account for variable vegetation cover in each grid cell, as follows:
- 210 $I_{c}(t) = f(t)(1 V(t)) + 4f(t)(V(t))$ (3) 211 $I_{a}(t) = Min[P(t), I_{c}(t)]$ (4)

where I_c is the infiltration capacity and V is the vegetation cover (between 0 and 1) in a model grid cell at time-step t. Values

used in the simulations for the parameters in equations 2-4 are provided in appendix Table A1.

214 3.3 Model setup

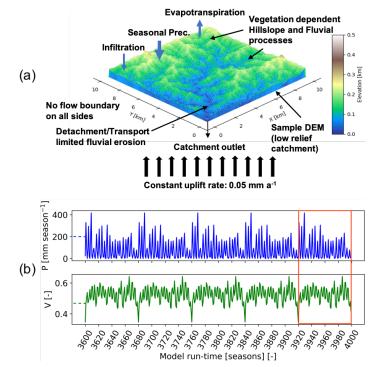
We applied the Landlab landscape evolution model (Hobley et al., 2017), combined with the SPACE 1.0 module of Shobe at al. (2017). The SPACE module allows coupled detachment-transport limited fluvial processes with simultaneous bedrock erosion and sediment entrainment/deposition. The Landlab-SPACE programs were modified for vegetation-dependent hillslope processes (Johnstone and Hilley, 2014) and vegetation-dependent overland flow and fluvial erosion using the approach described in Schmid et al. (2018) and Sharma et al. (2021). In addition, the geomorphic processes considered involve weathering and regolith production (Barnhart et al., 2019) and infiltration of surface water into soil (Rengers et al., 2016) based on the Green-Ampt method (Green and Ampt, 1911), and runoff modeling.

222 The model parameters (Table. A1) are calibrated to the distinct climate and ecological settings in the Chilean Coastal Cordillera 223 observations of Schaller et al., (2018). The model state parameters (i.e., erosion, diffusion, lithology, tectonic rock uplift rate, 224 etc.) in the simulations are adapted from Sharma et al., (2021). The model was simulated at a seasonal scale (time step of three months) from the Autumn of 2000 (01/03/2000) to the Summer of 2019 (28/02/2020). Simulations were conducted for a total 225 226 time of 1000 years with a time-step of 1 season (3 months) with 20 years (2000 - 2019) of observations in vegetation and precipitation. These 20-years of observations were repeated (looped) 50 times, to identify, and detrend, long-term transient 227 228 trends in catchment erosion rates due to potential differences in assumed erosional parameters such as the hillslope diffusivity or fluvial erodibility. The combined effect of temporally variable (at seasonal scale) precipitation and vegetation cover (also 229

230 spatially variable) on catchment-scale erosion rates was therefore the primary factors influencing predicted erosion rates.







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Figure 3. Schematic of the model geometry and seasonal precipitation and vegetation forcings used in this study. (a) Model setup representing sample DEM (low relief catchment) with no flow boundaries on all sides and a single catchment outlet. The model involves vegetation-dependent seasonal hillslope and fluvial processes and rainfallinfiltration-runoff modeling. (b) Seasonal precipitation and vegetation cover dataset (Mediterranean, La Campana, setting) for the last five iterations of model simulations. The results of highlighted iterations (after detrending for longterm transients) are analyzed in consecutive sections.

238 **3.4 Boundary and initial conditions**

The boundaries are closed (no flow) on all sides, with a single stream outlet at the point of minimum elevation at boundary nodes (Fig. 3). Initial sediment cover thickness is considered uniform across the model domain, and was approximated based on observations by Schaller et al., (2018) and Dal Bo et al., (2019). The sediment thickness used are 0.2 m in arid (AZ), 0.45 m in semi-arid (SG), 0.6 m in the Mediterranean (LC), and 0.7 m in humid temperate (NA) catchments. The initial rock uplift rate is kept constant throughout the entire model run as 0.05 mm yr⁻¹, adapted from a similar study (Sharma et al., 2021).

244 **3.5 Overview of simulations conducted**

The simulations were designed to identify the sensitivity of erosion rates to seasonal variations in either precipitation rates or vegetation cover, as well as the more realistic scenario of coupled seasonal variations in both vegetation cover and precipitation. We evaluated this sensitivity with a step-wise increase in model complexity. Three sets of simulations were designed for the four selected study areas, which are as follows,

- Scenario 1: Influence of constant (mean seasonal) precipitation with seasonal variations in vegetation cover
 catchment-scale erosion rates.
- Scenario 2: Influence of seasonal variation in precipitation and constant (mean seasonal) vegetation cover on
 catchment-scale erosion rates.





- Scenario 3: Influence of coupled seasonal variations in both precipitation and vegetation cover on catchment-scale
 erosion rates.
- 255 The results for scenarios 1 3 are illustrated in sections 4.1, 4.2, and 4.3, respectively.
- 256 **3.6 Detrending of results for long term transients**

Model simulations were conducted for 1,000 years using 20 years [March-2000 - Feb-2020] of observations in vegetation 257 cover, and precipitation and were repeated 50 times for a total simulation duration of 1000 years. Simulations presented here 258 259 were conducted on the present-day topography to allow for the application of observed time series of precipitation and vegetation change in different ecosystems and study areas. This choice of setting comes with the compromise that the erosional 260 parameters (e.g., diffusivity, erodibility, etc.) used in the model are likely not the same as those that led to the present-day 261 262 catchment topography. As a result, a long-term transient in erosion rates is expected as the model tries to reach an equilibrium with assumed erosional parameters. To correct for any long-term transients in erosion influencing our interpretations, we 263 264 conducted a linear detrending of the results to remove any long-term variations. Hence, the detrended model results for the last 20 years were analyzed and discussed in sections 4 and 5. 265

266 4 Results

- 267 In the following sections, we focus our analysis on the mean catchment erosion rates over seasonal (3 months) time scales (see
- Table. 1). In all scenarios, the rock uplift rate was kept constant at 0.05 mm yr⁻¹ following the approach of Sharma et al. (2021).
- 269 For simple representation, the results of the last five years of the last cycle of transient simulations starting from Autumn-2015
- to Summer-2019 are displayed in Fig. 4, 6, and 8 (after detrending, see section 3.6). The results for the entire time series
- 271 (Autumn-2000 Summer-2019) are available in the supplement (Fig. 1 3). The precipitation and erosion rates are shown
- with the units [mm season⁻¹].

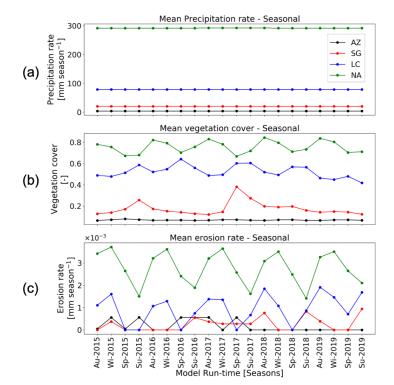
273 4.1 Scenario 1: Influence of constant precipitation and seasonal variations in vegetation cover on erosion rates

- In scenario 1, vegetation cover (MODIS NDVI from March 2000 to February 2020) fluctuates seasonally (Fig. 4b), and precipitation rates are kept constant at the seasonal mean (i.e., MAP divided by the number of seasons in a year) during the
- entire time-series (Fig. 4a) (Ziese et al., 2020). The range of seasonal vegetation cover variations (and mean seasonal
- precipitation rates) are observed as 0.06 0.08 [-] (3.92 mm season⁻¹), 0.1 0.4 [-] (20.16 mm season⁻¹), 0.35 0.65 [-] (79
- mm season⁻¹), and 0.5 0.85 [-] (292 mm season⁻¹) for the Pan de Azúcar, Santa Gracia, La Campana and Nahuelbuta study
- areas (Fig. 1), respectively.
- The predicted mean catchment seasonal erosion rates range between $0 6 \times 10^{-4}$ mm season⁻¹, $0 9.4 \times 10^{-4}$ mm season⁻¹ 1, $0 - 2.3 \times 10^{-3}$ mm season⁻¹, and $1.2 \times 10^{-3} - 4 \times 10^{-3}$ mm season⁻¹ in Pan de Azúcar, Santa Gracia, La Campana, and Nahuelbuta, respectively (Fig. 4c). The mean catchment seasonal erosion rates have an inverse linear relationship with seasonal vegetation cover for arid, semi-arid, and Mediterranean settings (Fig. 5). However, this relationship is positive in the humid-
- temperate setting, i.e., erosion increases with an increase in vegetation cover with a relatively lower gradient (3×10^{-3}) .
- 285 The maximum gradient between vegetation cover and erosion rates is observed in the Mediterranean region (La Campana,
- 286 gradient: -6×10^{-3}). The slopes in the vegetation cover erosion rate relationship (Fig. 5) represent the sensitivity of each
- 287 catchment to changes in seasonal vegetation cover, which indicates that the Mediterranean region (La Campana) is ~4.5 times
- more sensitive than the semi-arid region (Santa Gracia). Due to very low precipitation in the arid region (Pan de Azúcar), no significant range in erosion rates is observed (e.g., Pearson r: 0.17; p: >0.05). The results (Fig. 4 and 5) suggest a high sensitivity
- significant range in erosion rates is observed (e.g., Pearson r: 0.17; p: >0.05). The results (Fig. 4 and 5) suggest a high sensitivity in erosion in the Mediterranean setting to changes in seasonal vegetation cover (i.e., erosion rates decrease with an increase in





- 291 vegetation cover). The erosion rates are low (e.g., <0.005 mm season-1) due to low mean precipitation rates subjected to
- 292 infiltration and evapotranspiration.



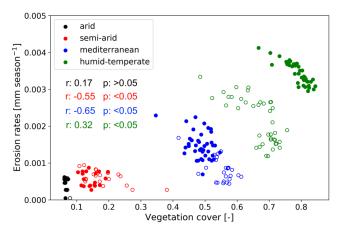
293

294 Figure 4. Results of simulations with constant seasonal precipitation and variable vegetation over last 5 years (Autumn-

295 2015 – Summer-2019) of last cycle of transient-state model run representing: (a) mean catchment seasonal precipitation

rates [mm season⁻¹], (b) mean catchment seasonal vegetation cover [-], and (c) mean catchment seasonal erosion rates

297 [mm season⁻¹].



298

299 Figure 5. Correlation of vegetation cover [-] and erosion rates [mm season⁻¹] obtained from simulations with constant

300 seasonal precipitation and variable vegetation over the last cycle of the transient state model run (Autumn-2000 -

301 Summer-2019). Hollow circles: dry season; filled circles: wet season. Each individual circle represents one predicted

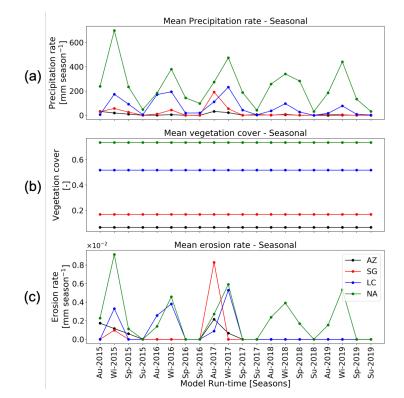
302 season within the timeseries.





4.2 Scenario 2: Influence of seasonal variations in precipitation and constant vegetation cover on erosion rates

- 304 In scenario 2, vegetation cover (MODIS NDVI from Mar-2000 Feb-2020) is kept constant at the mean seasonal vegetation
- 305 cover (Fig. 6b) and precipitation rates vary seasonally (Mar-2000 Feb-2020) (Fig. 6a). The range of seasonal precipitation
- rate variations are observed in the range of 0 32.42 mm season⁻¹, 0 191.66 mm season⁻¹, 0.03 417 mm season⁻¹, and 26 32.42 mm season⁻¹, 0 191.66 mm season⁻¹, 0 32.42 mm sea
- 307 987 mm season⁻¹ for Pan de Azúcar, Santa Gracia, La Campana and Nahuelbuta, respectively.



308

309 Figure 6. Results of simulations with variable seasonal precipitation and constant vegetation over last 5 years (Autumn-

2015 – Summer-2019) of last cycle of transient-state model run representing: (a) mean catchment seasonal precipitation
 rates [mm season⁻¹], (b) mean catchment seasonal vegetation cover [-], and (c) mean catchment seasonal erosion rates

The simulated mean catchment seasonal erosion rates are observed in the range of $0 - 2 \times 10^{-3}$ mm season⁻¹, $0 - 8.3 \times 10^{-3}$

314 mm season⁻¹, $0 - 1.37 \times 10^{-2}$ mm season⁻¹, and $0 - 1.3 \times 10^{-2}$ mm season⁻¹ in Pan de Azúcar, Santa Gracia, La Campana,

315 and Nahuelbuta, respectively (Fig. 6c). The mean catchment seasonal erosion rates are positively correlated with seasonal

316 precipitation rates (Fig. 7), with a maximum gradient in the arid region (AZ, gradient: $\sim 1.3 \times 10^{-4}$).

317 The gradients in the precipitation – erosion rate relationship (Fig. 7) indicate the sensitivity of each catchment to changes in

seasonal precipitation rates, such that the trid region (AZ) is ~2.7, ~3.2, and ~8 times more sensitive than semi-arid (SG),

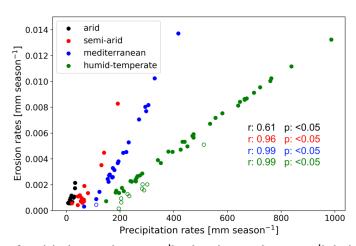
319 Mediterranean (LC) and humid-temperate region (NA), respectively. The results (Fig. 6 and 7) suggest a high sensitivity of

- 320 erosion in the arid and semi-arid settings to changes in seasonal precipitation rates (i.e., erosion increases at relatively higher
- 321 rates with an increase in precipitation). The erosion rates are higher than in scenario 1 (e.g., 0 0.014 mm season⁻¹) due to
- 322 higher precipitation rates in scenario 2.

^{312 [}mm season⁻¹].







323

Figure 7. Correlation of precipitation rates [mm season⁻¹] and erosion rates [mm season⁻¹] obtained from simulations with variable seasonal precipitation and constant vegetation over the last cycle of the transient state model run (Autumn-2000 – Summer-2019). Hollow circles: dry season; filled circles: wet season. Each individual circle represents one predicted season within the timeseries.

4.3 Scenario 3: Influence of coupled seasonal variations in both precipitation and vegetation cover on erosion rates

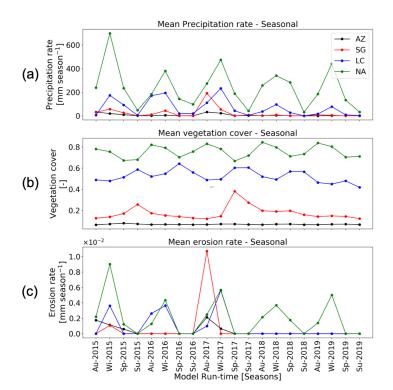
In this scenario, coupled variations in seasonal vegetation cover (MODIS NDVI from Mar-2000 – Feb-2020) (Fig. 8b) and precipitation rates are presented for the years 2000 - 2019 (Fig. 8a). The range of seasonal precipitation rates (and seasonal vegetation cover, V) variations are 0 - 32.42 mm season⁻¹ (V= 0.06 - 0.08 [-]), 0 - 191.66 mm season⁻¹ (0.1 - 0.38 [-]), 0.03- 417 mm season⁻¹ (0.35 - 0.65 [-]), and 26 - 987 mm season⁻¹ (0.5 - 0.85 [-]) for Pan de Azúcar, Santa Gracia, La Campana and Nahuelbuta, respectively.

The mean catchment seasonal erosion rates are observed in the range of $0 - 2 \times 10^{-3}$ mm season⁻¹, $0 - 1 \times 10^{-2}$ mm season⁻¹ 334 1 , $0 - 1.4 \times 10^{-2}$ mm season⁻¹, and $0 - 1.4 \times 10^{-2}$ mm season⁻¹ in Pan de Azúcar, Santa Gracia, La Campana, and 335 336 Nahuelbuta, respectively (Fig. 8c). Similar to scenario 2, mean catchment seasonal erosion rates are observed to be positively 337 correlated with seasonal precipitation rates (Fig. 9), with a maximum gradient in an arid region (AZ, gradient: $\sim 1.3 \times 10^{-4}$). 338 The slopes in the precipitation - erosion rate relationship (Fig. 9) represent the sensitivity of each catchment to coupled variations in seasonal precipitation rates and vegetation cover. The results (Fig. 8 and 9) indicate that the arid region (AZ) is 339 340 ~2.3, ~3, and ~8 times more sensitive than the semi-arid (SG), Mediterranean (LC), and humid-temperate region (NA), 341 respectively. The similarity in results obtained from scenarios 2 and 3 suggest the first-order control of seasonal precipitation changes on erosion rates (e.g., Pearson r > 0.6 for arid setting, Fig. 9a), with less significance to changing vegetation cover 342 343 (Pearson r < -0.19 for semi-arid setting, Fig. 9b).

344



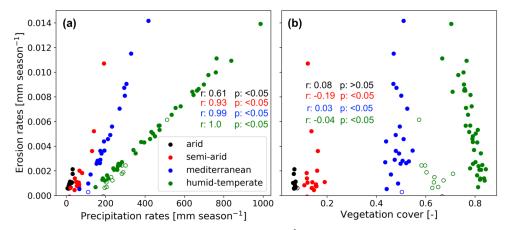




345

346Figure 8. Results of simulations with coupled variations in seasonal precipitation and vegetation over the last five years347(Autumn-2015 – Summer-2019) of the last cycle of transient-state model run representing: (a) mean catchment seasonal348precipitation rates [mm season⁻¹], (b) mean catchment seasonal vegetation cover [-], and (c) mean catchment seasonal

349 erosion rates [mm season⁻¹].



350

Figure 9. Correlation between (a) precipitation rates [mm season⁻¹] and (b) vegetation cover and erosion rates [mm season⁻¹] obtained from simulations with coupled variations in seasonal precipitation and vegetation over the last cycle

of the transient simulation (Autumn-2000 – Summer-2019). Hollow circles: dry season; filled circles: wet season. Each

- 354 individual circle represents one predicted season within the timeseries.
- 355





357 **5** Discussion

358 This section discusses the relationship between variations in seasonal precipitation and vegetation cover with erosion rates 359 (section 5.1). This is followed by a discussion of the effect of variable vegetation and precipitation rates on seasonal erosion 360 rates (section 5.2). Following this, we present the synthesis of catchment scale erosion rates variability over wet and dry 361 seasons (section 5.3). Finally, we compare our results with previously published studies (section 5.4) and discuss model 362 limitations (section 5.5).

363 5.1 Synthesis of the amplitude of change in erosion rates for model scenarios 1-3

364 The amplitude of change of mean catchment erosion rates [in percentage] varies at a seasonal scale (Fig. 10) between the study 365 areas. The amplitude of change in erosion rates to their respective mean values was estimated (Fig. 10) using the coefficient 366 of variation in percent (standard deviation divided by the mean of a dataset). The coefficient of variation is a statistical tool to 367 compare multiple variables free from scale effects, i.e., it is a dimensionless quantity (Brown, 1998). This comparison 368 represents the sensitivity of each catchment to changing seasonal weather for all three model scenarios (section 4.1 - 4.3).

In scenario 1, with seasonal variations in vegetation cover and constant seasonal precipitation (Fig. 10a), the amplitude of 369 370 change in erosion rates ranges between 6.5% in the humid-temperate setting and 36% in Mediterranean setting. The above

371 results support the findings of Zhang et al. (2019), which observed 20-30% of the total change in sediment yield with constant

372 precipitation and variable vegetation cover. The above study used the soil and water assessment tool (SWAT) based on NDVI

373 and climate parameters. In addition, a 5.5% change in amplitude in erosion rates is observed in the arid setting (Fig. 10a).

374 However, due to the weak correlation between vegetation cover and erosion rates (i.e., Pearson r: 0.17 and p>0.05, Fig. 5) and

- 375 negligible vegetation cover (V \leq 0.1), it is unclear if these changes in erosion rates are due to changes in vegetation cover
- 376 alone.

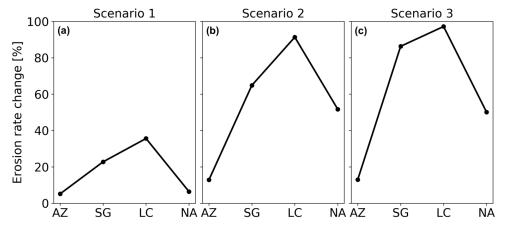






Figure 10. The amplitude of change in seasonal erosion rates (relative to their respective means) in (a) scenario 1: with 379 variable vegetation cover and constant precipitation rates, (b) scenario 2: with constant vegetation cover and variable 380 precipitation rates, and (c) scenario 3: with coupled variations in vegetation cover and precipitation rates.

381 In scenario 2, with constant vegetation cover and variable precipitation rates (Fig. 10b), the amplitude of change in erosion 382 rates ranges from 13% in the arid setting (AZ) to 52%, 65%, and 91% in humid-temperate (NA), semi-arid (SG) and 383 Mediterranean (LC) settings, respectively. A similar trend is observed in scenario 3 with coupled variations in vegetation cover 384 and precipitation rates (Fig. 10c), with the amplitude of change in erosion rates between 13% in the arid setting up to 50%, 385 86%, and 97% in the humid-temperate, semi-arid and Mediterranean settings, respectively. The magnitude of erosion rate changes is amplified in scenario 3, especially in the semi-arid setting (e.g., ~21% increase in the amplitude of change from 386





- 387 scenario 2 to scenario 3). This amplification could be owed to the 35% change in vegetation cover in the semi-arid setting (Fig.
- 388 8). Overall, these observations indicate a high sensitivity of erosion in semi-arid and Mediterranean environments comparedto arid and humid-temperate settings.
- 390 The pattern of erosion rate changes in scenarios 1-3 implies a predominant control of precipitation variations (rather than
- 391 vegetation cover change) on catchment erosion rates at a seasonal scale. This interpretation is consistent with previous
- 392 observational studies. For example, a plot-scale study by Gabarrón-Galeote et al. (2013) in the Mediterranean environment in
- 393 Belgium concluded that rainfall intensity was the main factor in determining the observed seasonal soil hydrological and
- solution erosive response. Another field study by Suescún et al. (2017) in the Columbian Andes highlighted the significant influence
- 395 of precipitation seasonality (over vegetation cover seasonality) on runoff and erosion rates. An observational catchment-scale
- 396 study in the semi-arid Chinese Loess Plateau by Wei et al. (2015) indicated that intra-annual precipitation variations were a
- 397 significant contributor to monthly runoff and sediment yield variations.

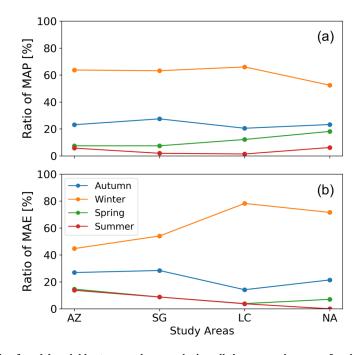
398 5.2 Synthesis of catchment erosion rates over wet and dry seasons

399 In this section, we discuss the ratio of MAP (Fig. 11a) and mean annual erosion (MAE) (Fig. 11b) during different seasons

400 (i.e., autumn - summer) in a year, averaged over the entire time series (2000 - 2019). This analysis is performed for the

401 simulation results of scenario 3 for different climate and ecological settings (i.e., arid to humid-temperate). We do this

- 402 specifically with scenario 3 results to capture the trends in erosion rates with coupled variations in model input (i.e.,
- 403 precipitation and vegetation cover).



404

Figure 11. The ratio of model variables to annual means during all the seasons in a year for simulation results from scenario 3 (coupled seasonal variations in precipitation and vegetation cover). The plots correspond to (a) ratio of MAP per season [%] and (b) ratio of MAE per season [%] of the mean values averaged over the last 20 years of the simulations (2000 - 2019). Each point represents the ratio for each ecological setting i.e., arid (Pan de Azúcar, AZ),

409 semi-arid (Santa Gracia, SG), Mediterranean (La Campana, LC), and humid-temperate settings (Nahuelbuta, NA).





410 The values for the ratio of MAP during different seasons (Fig. 11a) depicts winter (June-August) and summer (December-411 February) as the wettest and driest seasons of the year, respectively. For example, all study areas receive >50% and <6% of 412 MAP during winters and summers. The same is reflected in Fig. 11b with 45%, 55%, 78%, and 71% of MAE in the arid, semi-413 arid, Mediterranean, and humid-temperate settings, respectively, during winters. On the contrary, during summers the share of 414 MAE decreases from 14% in the arid setting to 1% in the humid-temperate setting. The Autumn (March-May) receives lower 415 precipitation amounts that range from 20-30% of MAP in the study areas. Arid and semi-arid settings experience a relatively 416 higher share of MAE (e.g., ~30%) than the Mediterranean and humid temperate settings (e.g., ~15-20%). The Spring season experiences relatively higher erosion rates despite a smaller share of MAP in arid and semi-arid settings. For example, the arid 417 and semi-arid settings experience 10-14% of the MAE for ~7% of MAP. At the same time, the Mediterranean and humid-418 419 temperate settings experience 5-7% of MAE for ~12-18% of MAP during Spring. Overall, we find that arid and semi-arid 420 settings experience <15% and ~50% of MAE during the wet (winter) and dry (summer) seasons. The above relationship is 421 amplified for the Mediterranean and humid-temperate settings with <5% and >70% of MAE occurring during wet and dry 422 seasons, respectively. The latter is in agreement with an observational study by Mosaffaie et al., (2015) in a Mediterranean 423 catchment in Iran. More specifically, Mosaffaie et al., (2015) used field observations from 2012-2013 to conclude that

424 maximum erosion rates (>70%) are observed during the wet season, which decreases in the dry season (<10%).

425 **5.3 Comparison to previous studies**

426 In this section, we relate the broad findings of this study to the previously published observational studies. In an observational 427 study in an agrarian drainage basin in the Belgian Loam Belt, Steegen et al., (2000) evaluated sediment transport over various 428 time scales (including seasonal). They observed lower sediment fluxes during the seasons with high vegetation cover. In 429 addition, an observational study by Zheng (2006) investigated the effect of vegetation changes on soil erosion in the Loess 430 Plateau, China, and concluded that soil erosion was significantly reduced (up to ~50%) after vegetation restoration. Another 431 observational study in semi-arid grasslands in Loess Plateau, China, by Hou et al., (2020) highlighted a considerable reduction 432 in erosion rates due to the development of richness and evenness of the plant community in the early to the mid wet season. 433 Our results from scenario 1 (seasonal variations in vegetation cover with constant precipitation rates) support the findings of 434 the above studies whereby a negative correlation (Pearson r: ~ -0.6 and p < 0.05) was found between vegetation cover and 435 erosion rates for the semi-arid and Mediterranean settings. More specifically, we found erosion rates decrease with an increase 436 in vegetation cover in Santa Gracia (semi-arid) and La Campana (Mediterranean) (see Fig. 5). However, a positive correlation 437 (Pearson r: ~ 0.3 and p< 0.05) is observed in the humid-temperate setting from dry season to wet season (see Fig. 5).

438 A catchment-scale observational study in Baspa Valley, NW Himalayas (Wulf et al., 2010), analyzed seasonal precipitation 439 gradients and their impact on fluvial erosion using weather station observations (1998 - 2007). The study observed a positive 440 correlation between precipitation and sediment yield variability, demonstrating the summer monsoon's first-order control on 441 erosion processes. An observational study by Wei et al., (2015) in Loess Plateau, China, evaluated erosion and sediment 442 transport under various vegetation types and precipitation variations. They found that significant changes in vegetation cover 443 might contribute to long-term soil dynamics. However, seasonal variations in runoff and sediment yield were mainly influenced 444 by rainfall seasonality. In comparison to the results of this study, we find the similarity in the patterns of erosion rates in 445 scenario 2 (variable precipitation and constant vegetation cover) and scenario 3 (coupled variations in precipitation and 446 vegetation) are consistent with the findings of Wei et al., (2015). For example, the amplitude of change in erosion rates (Fig. 447 10) in scenarios 2 and 3 differ by 0%, 6%, and -2% in the arid, Mediterranean, and humid-temperate settings, respectively. 448 However, this difference is enhanced in the semi-arid region (i.e., ~23%) due to a relatively high degree of variation (~25%) 449 in seasonal vegetation cover change.

- 450 Finally, an observational study in the Columbian Andes by Suescún et al., (2017) assessed the impact of seasonality on
- 451 vegetation cover and precipitation and found higher erosion rates in regions with steeper slopes. Another study by Chakrapani





- 452 (2005) emphasized the direct impact of local relief and channel slope on sediment yield in natural rivers. The broad findings
- 453 of the above studies agree with our results from scenarios 1-3, as we find higher erosion rates in the Mediterranean and humid-
- temperate regions with steeper topography (mean slope ~20 deg), which encounter high seasonality (and intensity) in
- 455 precipitation.

456 5.4 Model Limitations

The model setup used in this study was designed to quantify the sensitivity of erosion rates in different climate and ecological settings with variations in precipitation rates and vegetation cover at seasonal scales. We represent the degree of variations in

- erosion rates in terms of change amplitude (with respect to the mean) for different model scenarios (see sections 4.1 4.3).
- 460 This study was intended to introduce temporal downscaling (from millennial to seasonal time scales) to the approach of
- This study was interacted to introduce emporting downsearing (non-initialities searces) to the
- previous similar modeling studies (e.g., Schmid et al., 2018; Sharma et al., 2021).
 Our modeling approach used several simplifying assumptions that warrant discussion and potential investigation in future
- studies. For example, model results presented here successfully capture the major surface processes, including vegetation-
- 464 dependent erosion and infiltration, sediment transport, and surface runoff. However, groundwater flow is not considered in 465 the current study, and how the reentry of groundwater into streams over seasonal scales would influence downstream erosion.
- 466 The reason is that groundwater flow modeling includes a high amount of heterogeneity and anisotropy and requires much finer
- 467 grid sizes (<1m) and smaller time steps (in seconds to hours). Thus, due to the large grid-cell size (90 m), timescales (monthly),
- and high uncertainty in subsurface hydrologic parameters we were unable to evaluate the effects of groundwater flow on our
- results. Furthermore, this study assumed uniform lithological and hydrological parameters (e.g., vertical hydraulic conductivity, initial soil moisture, evapotranspiration, erodibility, etc.) over the entire catchment. As said earlier, these
- 471 properties are subjected to a high level of uncertainty and heterogeneity, the best fitting parameters, based on previously
- 472 published literature (e.g., Schaller et al., 2018; Bernhard et al., 2018; Schmid et al., 2018; Sharma et al., 2021) are used for the
- 473 model simulations. However, the heterogeneity in vegetation cover and related soil-water infiltration per grid cell is used in
- 474 this study. For the heterogeneity in vegetation cover, we use MODIS-derived NDVI as a proxy of vegetation cover. According
- 475 to Garatuza-Payán et al. (2005), NDVI is assumed as an effective tool for estimating seasonal changes in vegetation cover
- 476 density. However, the spatial resolution (250 m) of the NDVI dataset is lower than that of the SRTM DEM (90 m) used in the
- study. Nevertheless, the difference in spatial resolution of vegetation cover and topography might introduce ambiguity in themodel results.
- A final limitation stems from several generalized model parameters (e.g., rock uplift rate, erodibility, diffusivity, etc.) applied to the SRTM DEM (as initial topography). We did this to capture the effects of seasonality in precipitation and vegetation cover in modern times (2000 - 2019). However, the current topography might not have evolved with the same tectonic and lithological parameters. To address this limitation, we simulated the model for 50 iterations and detrended the model results to remove those transient effects (see section 3.6). This limitation can be handled in future studies by parameterizing the model to the current topography using stochastic (e.g., Bayesian) techniques (e.g., Stephenson et al., 2006; Avdeev et al., 2011). As this study was aimed to capture the control of seasonal precipitation and (or) vegetation changes on the relative variability of
- 486 erosion rates, the above limitation may not pose a problem in the model results.

487 6 Summary and Conclusions

In this study, we applied a landscape evolution model to quantify the impact of seasonal variations in precipitation and vegetation on catchment averaged erosion rates. We performed this in regions with varied climate and ecology including: arid, semi-arid, Mediterranean, and humid-temperate settings. Three sets of simulations were designed to model erosion rates for (a) scenario 1: constant precipitation and variable vegetation cover, (b) scenario 2: variable precipitation and constant





492 vegetation cover, and (c) scenario 3: coupled variations in precipitation and vegetation cover. The main conclusions derived493 from this study are as follows:

- Scenario 1, with variable vegetation cover and constant precipitation (Fig. 4), resulted in small variations in seasonal erosion rates (<0.02 mm yr⁻¹) in comparison to the other scenarios. The amplitude of change in seasonal erosion rates (relative to the mean) is the smallest in humid-temperate setting and maximum in the Mediterranean setting (Fig. 10a). For example, it ranges from 6.5% in humid-temperate setting (Nahuelbuta) to 23% and 36% in semi-arid (Santa Gracia) and Mediterranean settings (La Campana), respectively.
- Scenario 2, with constant vegetation cover and variable precipitation (Fig. 6), results in relatively higher seasonal erosion rates (<0.06 mm yr⁻¹) in comparison to scenario 1. The amplitude of change in seasonal erosion rates (relative to the mean) is smallest in the arid setting and the largest in the Mediterranean setting (Fig. 10b). For example, it ranges from 13% in the arid setting (Pan de Azúcar) to 52%, 65%, and 91% in humid-temperate (Nahuelbuta), semi-arid (Santa Gracia), and Mediterranean settings (La Campana), respectively.
- Scenario 3, with coupled variations in vegetation cover and precipitation (Fig. 8), results in similar seasonal erosion rates (<0.06 mm yr⁻¹) to scenario 2. Similarly, the amplitude of change in seasonal erosion rates (relative to the mean) is the smallest in the arid setting and the largest in the Mediterranean setting (Fig. 10c). For example, it ranges from 13% in the arid setting (Pan de Azúcar) to 50%, 86%, and 97% in humid-temperate (Nahuelbuta), semi-arid (Santa Gracia), and Mediterranean settings (La Campana), respectively. A significant increase (from scenario 2) in variation in erosion rates (~21%) is owed to the ~25% variation in vegetation cover in semi-arid settings.
- All study areas experience maximum and minimum erosion during wet and dry seasons, respectively (Fig. 11b).
 However, the difference (in maximum and minimum) is amplified from the arid (~30%) to the Mediterranean and humid-temperate settings (~70-75%). This is owed to the range of amplitude of precipitation rate change (Fig. 7) increasing from the arid (e.g., ~9 mm) to humid-temperate settings (e.g., ~543 mm) in wet and dry seasons.

514 Finally, this study was motivated by testing the hypotheses that (1) if precipitation variations primarily influence seasonal 515 erosion, then the influence of seasonal vegetation cover changes would be less significant, and (2) catchment erosion in drier 516 settings is more sensitive to seasonality in precipitation and vegetation, than wetter settings. With respect to hypothesis 1, we 517 found that seasonal precipitation variations primarily drive catchment erosion and the effects of vegetation cover variations 518 are secondary. Results presented here (Fig. 10b) support this interpretation with significantly high amplitude of change in 519 erosion rates (with respect to means) ranging from 13 to 91% for scenario with constant vegetation cover and seasonal 520 precipitation variations. However, the effect of seasonal vegetation cover changes is also significant (Fig. 10a) ranging between 521 5 - 36%. Hence, the first hypothesis is partially confirmed. Concerning hypothesis 2, we found that seasonal changes in 522 catchment erosion are more pronounced in semi-arid and Mediterranean settings and less in arid and humid temperate settings. 523 This interpretation is supported by Fig. 10c, with significantly high amplitude of change in catchment erosion in semi-arid 524 (~86%) and Mediterranean (~97%) settings with relatively lower changes in humid temperate (~50%) and arid (~13%) settings, 525 partially confirming the hypothesis.

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- 530





531 Appendix

532 Table A1. Input parameters with corresponding units for the landscape evolution model

Model Parameters	Values	
Grid spacing (dx)	90 m	
Model runtime (totalT)	1000 years (2000-2019 repeated over 50 times)	
time-step (dt)	1 season (3 months)	
Rock uplift rate (U) ¹	1.25 x 10 ⁻⁵ [m season ⁻¹] (or 0.05 [mm a ⁻¹])	
Initial sediment thickness (H_initial) ²	20 (AZ), 0.45 (SG), 0.6 (LC), 0.7 (NA) [cm]	
Bedrock erodibility (Kr) ¹	2 x 10 ⁻⁹ [m ⁻¹]	
Sediment erodibility (Ks)1	$2 \ge 10^{-8} [m^{-1}]$	
Reach scale bedrock roughness (H*) ¹	1 [m]	
Porosity $(\varphi)^1$	0.2 [-]	
Fraction of fine sediments (Ff) ¹	0.2 [-]	
Effective terminal settling velocity (Vs)1	2.5 [m season ⁻¹]	
m, n ¹	0.6, 1 [-]	
Bedrock erosion threshold stream power $(\omega_cr)^1$	1.25 x 10 ⁻⁵ [m season ⁻¹]	
Sed. entr. threshold stream power $(\omega_cs)^1$	$1.25 \ge 10^{-6} [m \text{ season}^{-1}]$	
Bare soil diffusivity (K _b) ¹	2.5 x 10 ⁻⁴ [m ² season ⁻¹]	
Exponential decay coefficient (a)1	0.3 [-]	
Critical channel formation area (A _{crit}) ³	1 x 10 ⁶ [m ²]	
Reference vegetation cover $(V_r)^3$	1 (100%)	
Manning's number for bare soil $(n_s)^3$	0.01 [-]	
Manning's number for ref. vegetation $(n_v)^3$	0.6 [-]	
Sacling factor for vegetation influence (w) ³	1 [-]	
Soil bulk density (B) ⁴	1300 (AZ), 1500 (SG), 1300 (LC), 800 (NA) [kg m-3]	
Soil type ⁴	sandy loam (AZ, SG, LC); sandy clay loam (NA)	
Initial soil moisture (s) ⁴	0.058 (AZ), 0.02 (SG), 0.053 (LC), 0.15 (NA) [m ³ m ⁻³]	

533

¹ Sharma et al. (2021), ² Schaller et al. (2018), ³ Schmid et al. (2018), ⁴ Bernhard et al. (2018).

535 Code and data availability

536 The code and data used in this study are freely available upon request.

537 Author contributions

- 538 HS and TAE designed the initial model setup and simulation programs. HS and TAE conducted model modifications,
- simulation runs, and analysis. HS prepared the paper with contributions from TAE.

540 Competing interests

541 The authors declare that they have no conflict of interest.

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