

Short communication: Concentrated impacts by tree canopy drips - hotspots of soil erosion in forests

Ayumi Katayama¹, Kazuki Nanko², Seonghun Jeong³, Tomonori Kume¹, Yoshinori Shinohara⁴, Steffen Seitz⁵

¹Shiiba research forest, Kyushu University, Miyazaki, 8830402, Japan

²Department of Disaster Prevention, Forestry and Forest Products Research Institute, Tsukuba, 3058687, Japan

³Department of Forest Environmental Resources, Gyeongsang National University, (Institute of Agriculture & Life Science), Jinju, 52828, Republic of Korea

⁴Faculty of Agriculture, University of Miyazaki, Miyazaki, 8892155, Japan

⁵Department of Geosciences, University of Tübingen, Tübingen, 72074, Germany

Correspondence to: Ayumi Katayama (ayumi.katayama0920@gmail.com)

Abstract.

The degradation of ground vegetation cover caused by large grazing herbivores frequently results in enhanced erosion rates in forest ecosystems. Splash erosion can be caused by drop impacts with a high throughfall kinetic energy (TKE) from the tree canopy. Notably bigger canopy drips from structurally-mediated woody surface points appear to induce an even higher TKE and generate concentrated impact locations causing severe focus points of soil erosion. However, TKE at these locations has rarely been reported. This pilot study investigated the intensity of TKE at a concentrated impact location and compared it with general TKE locations under the canopy and freefall kinetic energy (FKE) outside the forest. We measured precipitation, TKE and FKE using splash cups at seven locations under Japanese beech trees and five locations outside the forest during the leafless and leafed seasons in a deciduous broadleaved forest in Japan. The TKE at the concentrated impact location was 15.2 and 49.7 times higher than that at the general locations under the beech and FKE, respectively. This study confirmed that canopy drip from woody surfaces could be a hotspot of soil erosion in temperate forest ecosystems. Throughfall precipitation at the concentrated impact location was 11.4 and 8.1 times higher than that at general locations and freefall, respectively. TKE per 1 mm precipitation (unit TKE) at the concentrated impact location ($39.2 \pm 23.7 \text{ J m}^{-2} \text{ mm}^{-1}$) was much higher than that at general locations ($22.0 \pm 12.7 \text{ J m}^{-2} \text{ mm}^{-1}$) and unit FKE ($4.5 \pm 3.5 \text{ J m}^{-2} \text{ mm}^{-1}$). Unit TKE in the leafless season was significantly lower than in the leafed season because of fewer redistribution of canopy drips induced only by woody tissue. Nevertheless, unit TKE at the concentrated impact location in the leafless season ($36.4 \text{ J m}^{-2} \text{ mm}^{-1}$) was still higher than at general locations in the leafed season. These results show that potentially high rates of sediment detachment can be induced not only by throughfall precipitation, but also by larger throughfall drop size distributions at the concentrated impact location, even in the leafless season. [Further studies with higher replication building on this first report are necessary to investigate how many of these concentrated impact locations may occur on average on different tree species to better assess the extent of the erosion risk under forests.](#)

40 1. Introduction

41 Soil conservation is an important environmental challenge of the 21st century as soils are the foundation
42 of life and a reservoir for water, carbon, and nutrients (Lal, 2014). The threat to soil composition is
43 evident worldwide, especially in areas with regularly recurring extreme climatic events such as heavy
44 rainfall (Borrelli et al., 2020). Soil erosion rates induced by water are mainly determined by rainfall
45 patterns such as raindrop kinetic energy and ground cover by vegetation (Seitz et al., 2017). Severe soil
46 erosion events are rare in forest ecosystems because the general abundant ground cover owing to
47 understory vegetation or plant litter (Miura et al., 2003; Holz et al., 2015). Therefore, forest can be seen
48 as one of the most effective land use types to mitigate soil losses (Pimentel and Burgess, 2013). However,
49 disturbance of forest vegetation may lead to significant punctual (Gall et al., 2022; Geißler et al., 2010)
50 and areal (Safari et al., 2016; Seitz et al., 2016; Zemke et al., 2016) erosion events that can by far exceed
51 sustainable erosion rates (Deng et al., 2023). Important examples have been described globally such as in
52 Hungary (Misik and Kárász, 2022) and China (Yao et al., 2019). Especially in Japan, understory
53 vegetation in forests is regularly damaged by grading sika deer (*Cervus nippon*) (Murata et al., 2009,
54 Takatsuki, 2009). The degradation of protective vegetation layers frequently results in enhanced splash
55 erosion through direct raindrop impacts and increased surface runoff with significant erosion potential
56 (Shinohara et al., 2018; Song et al., 2019).

57 Throughfall kinetic energy (TKE, in J m^{-2}) can be determined from drop size and velocity in
58 addition to the precipitation amount. TKE has partly shown to be higher than freefall kinetic energy
59 (FKE) outside vegetation layers. This phenomenon is attributed to the capacity of the forest canopy to
60 generate large new canopy drips after the first interception, which depends on the species (Chapman,
61 1948; Nanko et al., 2015). Canopy drip can contribute to more than half of the total throughfall volume
62 from leafed canopies (Levia et al., 2019). In canopy water flow, the lateral redistribution plays an
63 important role in creating local concentration of throughfall (Keim and Link, 2018). Subsequently, lateral
64 canopy water flow paths ending at structurally-mediated woody surface drip points, such as irregular
65 rough points and branch concavities, accumulate more water volume transported down the branch with a
66 longer residence time and then generate larger diameter drops in greater volumes (Nanko et al., 2022)
67 than foliar surfaces (Levia et al., 2019; Nanko et al., 2016; Nanko et al., 2022). Notably bigger canopy
68 drips can have higher TKE and therefore, generate concentrated impact locations potentially causing
69 severe soil erosion. However, the TKE at these concentrated impact locations and the subsequent splash
70 erosion potential have rarely been described in the literature and not yet been quantified.

71 TKE is linearly correlated with throughfall precipitation in monolayer coniferous forests
72 (Shinohara et al., 2018). The slope of the relationship between throughfall precipitation and TKE is
73 known as unit TKE, that is, TKE per 1 mm precipitation. The unit TKE differs with canopy species and
74 architecture, and rainfall intensity (Nanko 2013; Nanko et al., 2015; Liu et al., 2022). Throughfall from
75 woody surface drip points consist of larger canopy drips, suggesting that the unit TKE at such
76 concentrated impact locations is different from that at other general locations. Furthermore, this
77 relationship might also differ between the leafed and leafless seasons, owing to the difference in the

78 distribution of drops of different sizes (Levia et al., 2017). Thus, TKE can considerably affect soil erosion
79 rates also in the leafless season when the contribution of drip points to the total throughfall precipitation
80 becomes dominant (Levia et al., 2019). Therefore, knowledge of the significance of TKE at concentrated
81 impact locations and seasonal changes in TKE in response to leaf status is vital for understanding soil
82 erosion risk in forests with degraded ground cover.

83 This [pilot](#) study investigated TKE under broadleaved trees in Shiiba Research Forest, Kyushu,
84 Japan, which is a substantially disturbed and eroded forest ecosystem caused by deer grazing. A special
85 focus of this [studywork](#) is given on unusual high energy levels induced by structurally-mediated woody
86 surface drip points which partly occurred during the measurement campaign to estimate throughfall
87 erosivity. We quantified the TKE intensity at the concentrated impact location. We hypothesized that: (1)
88 unit TKE at the concentrated impact location is higher than that at general locations inducing elevated
89 splash erosion, and (2) the relationship between throughfall precipitation and TKE differs with the leaf
90 status of trees.

91

92 **2. Materials and methods**

93 **2.1 Study site**

94 This study was conducted in Shiiba Research Forest, Kyushu, South Japan [32°40'N, 131°17'E, 1030 m
95 a.s.l.]. The study site includes a mixed forest with evergreen coniferous trees and deciduous broadleaved
96 trees. The mean annual temperature and precipitation were 10.8°C and 3278 mm, respectively, as
97 measured at a meteorological station located 3 km from the study site at 1180 m a.s.l. Monthly
98 precipitation amount in March, April, August, September and August of 2021 were 162, 133.5, 958.5,
99 170 and 41.5 mm at the University Forest office, situated 4k m away from the study site [600 m a.s.l.]
100 Formerly, this area was characterized by a dense understory vegetation comprised primarily of bamboo
101 (*Sasa borealis* [Hack.] Makino & Shibata). However, this understory vegetation has mostly disappeared
102 since around the year 2000, coinciding with a documented rise in the Sika deer population. Currently,
103 there is no intact understory vegetation in most areas of the research forest (Kawakami et al., 2020).
104 Therefore, distinct erosion forms and root exposure can be widely observed (Katayama et al., 2023) and
105 soil degradation has been identified a major challenge for forest services (Abe et al., 2022).

106

107 **2.2 Throughfall kinetic energy**

108 TKE was determined using splash cups (Shinohara et al., 2018; Scholten et al., 2011). The splash cups are
109 filled with a standardized sand and weighed dry before deployment in the field. Subsequently, raindrops
110 hit the sand surface and detached sand is partly splashed away from the cup. The loss of sand (LoS, g m⁻²)
111 was measured by back weighing remaining dried sand volumes and subtracting the amount from the
112 initial amount. TKE can be estimated from the relationship between KE and LoS using a linear function
113 (TKE = 14.55 × LoS, Scholten et al., 2011). This method has proven to be reliable and cost efficient with
114 a high number of replications (Geißler et al., 2010) and is suitable to evaluate spatial variation in TKE
115 (Shinohara et al., 2018). We used the splash cups with the diameter of 5.0 cm, height of 5.1 cm and the

116 volume of 100cc. These are slightly larger than those reported by Scholten et al., 2011 (4.6 of diameter
117 and 3.6 cm of height, respectively), but accurately estimated TKE by using a linear equation (Shinohara et
118 al. 2018).

119 The LoS was measured during each of the five rainfall events in the leafless (March to April)
120 and leafed (August to September) seasons of 2021. Seven splash cups were installed under the canopy of
121 two *Fagus crenata* trees for TKE (Fig. 2). One position was chosen at a possible concentrated drip
122 location formed by structurally-mediated wood surface, and where more throughfall precipitation was
123 visually observed during rainfall events. Six more splash cup at different positions under the canopy were
124 installed to measure TKE at general locations. Five splash cups were further installed outside the forest to
125 measure FKE where were 40 m apart from the locations under the canopy. A storage-type bottle with a
126 funnel (diameter: 9.0 cm) was installed next to each splash cups to measure precipitation. Precipitation
127 was measured at the same time with TKE measurement. The distance between the splash cup and
128 precipitation collector was about 5 cm.

129 At the concentrated impact location, the collection of LoS and throughfall precipitation failed for
130 some very strong rainfall events during the leafed period. We obtained data of 10 events at the general
131 locations (Table 1), but TKE and throughfall precipitation at the impact location were obtained only in
132 seven and six events. Thus, the relationship between TKE and freefall precipitation ($TKE = 237.1 \times$
133 freefall precipitation, $R^2 = 0.92$) was established using the data obtained in seven events whereas the
134 relationship between throughfall precipitation and freefall precipitation -(throughfall precipitation = 8.23
135 \times freefall precipitation, $R^2 = 0.97$) was established using the data obtained the six events.

136

137 **2.3 Tree traits**

138 Diameter at breast height of the two selected beech trees was 46.0 cm and 46.1 cm, and tree height was
139 21.1 m and 18.0 m, respectively. LAI was determined using a single reflex camera system with fish eye
140 lens (THETA SC; Ricoh Co. Ltd., Tokyo, Japan) and software (a Gap Light Analyzer ver. 2.0, Frazer et
141 al., 2022) was 4.5 and 0.9 at the concentrated impact location in the leafed and leafless season,
142 respectively. LAI at general locations ranged from 1.7 to 4.9 with a mean of 3.3 om the leafed season and
143 from 0.1 to 0.6 with a mean of 0.3 in the leafless season. Branch height at the concentrated impact
144 location was 9.1 m and ranged from 6.5 m to 13.5 m at the six splash cup positions with an average of 9.1
145 m. Average leaf area and leaf mass per area obtained from beech leaves in our study forest were 10.5 cm^2
146 and 84.7 g m^{-2} , respectively. The bark of the beech was smooth; however there was moss cover in some
147 places along the stem and epiphytic moss at the base of the branch, from which considerable amounts of
148 water dropped to the ground.

149

150 **2.4 Statistical analysis**

151 The significant difference in the slopes of the relationships of throughfall precipitation with TKE between
152 concentrated impact location and general locations was examined using ANCOVA ($P < 0.05$). The
153 significant difference in slopes in the relationships between the leafed and leafless seasons was examined

154 separately for impact and general locations (ANCOVA, $P < 0.05$). In these analyses, TKE data which was
155 not measured in the three rainfall events were excluded. Intercepts were set to zero for the models. All
156 statistical analyses were performed using *R* ver. 3.6.2 (*R* Core Team, 2019).

157

158 **3. Results and Discussion**

159 **3.1 Effect of structurally designed high energy points on TKE**

160 Considerably high TKE was observed at the concentrated impact location under the beech (Fig. 1). This
161 location received a focused number of canopy drips from an overlying structurally-mediated woody
162 surface drop point (Supplemental Video). Average \pm S.D. of TKE at the concentrated impact location
163 ($9142 \pm 5522 \text{ J m}^{-2}$) for all seasons was 15.2 times higher than at general locations under the beech ($601 \pm$
164 495 J m^{-2}) and 49.7 times higher than FKE ($184 \pm 195 \text{ J m}^{-2}$, Table 1). The average of throughfall
165 precipitation at the concentrated impact location ($324 \pm 227 \text{ mm}$) was 11.4 times higher than that at
166 general locations under beech ($29 \pm 16 \text{ mm}$) and 8.1 times higher than that from freefall precipitation (40
167 $\pm 26 \text{ mm}$).

168 Across all rainfall events, TKE significantly increased with throughfall precipitation at both the
169 concentrated impact location and general locations regardless of canopy leaf conditions (Fig. 3). It could
170 be shown that TKE at the concentrated impact location was higher than at general locations with a
171 significant difference in the relationships between TKE and throughfall precipitation (Fig. 3). Thus, the
172 first hypothesis can be confirmed. Furthermore, the branch height at the concentrated impact location was
173 comparable to average of branch height at other general drip points, indicating that higher unit TKE was
174 mostly induced by bigger drop sizes. Note that the unit TKE is determined from raindrop size
175 distributions and canopy height when the canopy height is less than the height for the rain-drop terminal
176 velocity (Shinohara et al., 2018). Previous study showed that most canopy drips did not reach to the
177 terminal velocity where the mean first living branch height was 7.9 m (Nanko et al., 2008). Raindrops
178 with diameters $>3 \text{ mm}$ need at least 12 m fall distance to gain terminal velocity (Wang and
179 Pruppacher, 1977). Although the branch height could be one of factors determining TKE in the present
180 study, considerable higher TKE at the impact location was not caused by the height because of the
181 comparable branch height. Thus, the TKE at the concentrated impact locations originating from woody
182 surface was induced by both high throughfall precipitation and big drop size, which is an important cause
183 of splash erosion and might be considered as an underestimated hot spot of sediment translocation.

184

185 **3.2 Effects of leaf status**

186 In the leafed season, the event-scale average TKE at the concentrated impact location was 12.5 times
187 higher than that at general locations under the beech tree and 61.5 times higher than FKE (Table1). The
188 event-scale mean throughfall precipitation at the concentrated impact location was 12.2 times higher than
189 at general locations and 8.1 times higher than freefall precipitation. In the leafless season, the average
190 TKE at the concentrated impact location was 23.6 times higher than that at general locations and 37.6
191 times higher than FKE, whereas mean throughfall was 10.3 times higher at general locations and 8.2

192 times higher than freefall precipitation. These results suggest that the splash erosion risk at the impact
193 location remained high in the leafless season although the risk was lower than that in the leafed season.
194 The ratios of throughfall precipitation at the concentrated impact location and general locations compared
195 to freefall precipitation were 8.1 and 0.71, respectively, suggesting that throughfall precipitation
196 decreased with canopy interception whereas the identified hotspot of throughfall selectively increased it.
197 The slopes of the relationships between TKE and throughfall precipitation at the concentrated impact
198 location and general locations were higher in the leafed season than in the leafless season (ANCOVA, $P <$
199 0.01). Therefore, we can conclude that unit TKE strongly increases with the presence of leaves and
200 potential splash erosion is higher during the leafed period. However, unit TKE at the concentrated impact
201 location in the leafless season ($36.4 \text{ J m}^{-2} \text{ mm}^{-1}$) was still higher than at general locations in the leafed
202 season ($32.1 \pm 10.3 \text{ J m}^{-2} \text{ mm}^{-1}$). This suggests high splash erosion risk at the concentrated impact
203 location even in the leafless season. In summary, leaf status has been shown to generate a distinct impact
204 and differential effects; therefore, the second hypothesis can be accepted.

205 Additionally, the differences between TKE and FKE as well as throughfall and freefall
206 precipitation appeared to be less pronounced in the leafless season. Levia et al., (2019) showed that
207 canopy drips under broadleaved trees accounted for 69% of the total throughfall precipitation in the leafed
208 phenophase, compared to 8% in the leafless phenophase. Most of the throughfall at general locations
209 under leafless trees were freefall. The soil erosion risk is lower during the leafless season than during the
210 leafed season except for the concentrated drop impact locations.

211

212 **3.3 Implication and uncertainty**

213 This study remarked notably high TKE under investigated beech trees. Mean unit FKE was reported by
214 van Dijk et al., (2002) as 14.2, 18.6, 26.5, and $28.1 \text{ J m}^{-2} \text{ mm}^{-1}$ with rainfall rates of 1, 10, 50, and 100
215 mm h^{-1} , respectively. The maximum measured unit FKE was $28.3 \text{ J m}^{-2} \text{ mm}^{-1}$. As for throughfall, unit
216 TKE reported in previous studies ranged from 16.4 to $28.1 \text{ J m}^{-2} \text{ mm}^{-1}$ in Japan (Nanko, 2013), Hawaii
217 (Nanko et al., 2015) and Thailand (Nanko et al., 2020). The unit TKE at the concentrated impact location
218 in the present study was much higher than these previously reported values, suggesting that higher
219 throughfall precipitation and larger throughfall drop size distributions can result in an increased risk of
220 soil erosion. Furthermore, unit TKE for general locations in the present study was also higher than in
221 previously measured Japanese cypress plantations with 16.4 - $21.0 \text{ J m}^{-2} \text{ mm}^{-1}$ (Nanko, 2013). The median
222 volume drop size of canopy drip from leaves was 4.7 mm in Japanese cypress but 5.2 mm in beech
223 (Nanko et al., 2013). This difference was caused by various leaf traits such as leaf area, leaf shape, and
224 leaf surface water repellency (Levia et al., 2017). Thus, TKE generation is strongly species specific and
225 TKE under beech trees may be higher than that under other tree species.

226 Finally, although considerable higher TKE at the concentrated impact location was measured
227 using splash cups, it should be noted that TKE at the concentrated impact location in the present study
228 may have been underestimated due to the rim effect related to the splash cup measuring system. There is
229 some uncertainty in the estimated TKE if sand particles are starting to hit the cup wall instead of flying

230 out. This phenomenon occurred particularly at the concentrated impact location. Thus, TKE at the
231 concentrated impact location may be even higher than that reported in the present study.

232 This pilot study showed a possible increase of soil erosion risk under the forest canopy using a
233 limited dataset obtained at one impact location. TKE and throughfall precipitation can considerably vary
234 with tree traits and thus, it is necessary to clarify spatial variations in TKE and throughfall and factors
235 determining variations such as tree traits and leaf status based on random sampling with higher replication
236 in further studies. Regarding the erosion potential, it is furthermore of high importance to investigate how
237 many of these concentrated impact locations may occur on average on different tree species to better
238 assess the general extent of the erosion risk

240 4. Conclusions

241 In this paper, we report the results of a pilot splash cup experiment conducted to investigate potential
242 erosion from one high energy water release points under the canopy in a disturbed Japanese forest
243 environment. ~~Extremely high~~Strongly enhanced TKE was observed from structurally-mediated woody
244 surface points under beech (*Fagus crenata*), which was approximately 15 times higher than that at
245 general locations and approximately 50 times higher than FKE. The higher kinetic energy was caused by
246 both higher throughfall precipitation and higher unit kinetic energy. These results underline the evidence
247 of high soil erosion risk in forested areas owing to particular tree traits and show that this risk can
248 significantly exceed the previously known dimensions at specific points under the tree canopy. Moreover,
249 unit TKE at high-energy and general locations was reduced in the leafless season, but unit TKE in the
250 leafless season was still higher at the concentrated impact location than at general locations in the leafed
251 season. This result points to a potentially enhanced soil erosion risk even outside the growing season if
252 concentrated impact locations with high kinetic energies occur in larger numbers on trees. ~~Furthermore, it~~
253 ~~is usually higher precipitation in the summertime in Japan because of rainy and typhoon season.~~
254 Precipitation amount is the most important factor determining soil erosion risk and higher precipitation
255 will also result in severe erosion risk in the leafed season. Further research is necessary to verify the
256 results, expand them to include other tree species and forest ecosystems and to shed ~~more~~ light on the
257 mechanistic effects of distinct plant characteristics building up on this pilot study. In this context, it
258 should also be investigated how many of these concentrated impact locations may occur on average on
259 different tree species to better assess the extent of ~~the soil~~ erosion. ~~risk because this study just showed~~
260 ~~considerable high soil erosion risk using dataset obtained at only one impact location.~~ This becomes
261 particularly important when the protective soil cover layer with the understory or leaf litter is disturbed or
262 removed. Therefore, future studies examining soil erosion rates under forests need to considerate both
263 changes in TKE through plant traits and variations in ground cover. Finally, as this pilot study showed
264 considerable high soil erosion risk using a dataset obtained at one single impact location, more
265 measurements under different trees with higher replication are needed.

267 Data availability

268 All raw data is provided in the supplement material.

269 **Video supplement**

270 <https://doi.org/10.5446/61199>

271 **Author contribution**

272 AK, KN and SS designed the experiment, AK, YS, TK and SJ carried it out. AK, KN and SS prepared the
273 manuscript with contributions from all co-authors.

274 **Competing interests**

275 The authors declare that they have no conflict of interest.

276 **Acknowledgments**

277 We thank the technical staff of Shiiba Research Forest who helped with the preparation and establishment
278 of measurements. We also thank Kyushu University Fund which allowed us to meet in Shiiba.

279 **Financial support**

280 This study was financially supported by JSPS KAKENHI Grant Number 22H03793 and JSPS
281 Postdoctoral Fellowship for Research in Japan (Short-term) Grant Number PE21018.

282

283 **References**

- 284 Abe, H., Fu, D., Kume, T., and Katayama, A.: Exposure of tree roots and its control factors in a mixed
285 temperate forest with no understory vegetation, *Bull. Kyushu Univ For.*, 103, 13–20,
286 <https://doi.org/10.15017/4776829> (Japanese with English summary), 2022
- 287 Borrelli, P., Robinson, D. A., Panagos, P., Lugato, E., Yang, J. E., Alewell, C., Wuepper, D.,
288 Montanarella, L., and Ballabio, C.: Land use and climate change impacts on global soil erosion by
289 water (2015-2070), *Proc. Natl. Acad. Sci.*, 117, 21994–22001,
290 <https://doi.org/10.1073/pnas.2001403117>, 2020.
- 291 Chapman, G.: Size of raindrops and their striking force at the soil surface in a red pine plantation. *Eos*,
292 *Transactions American Geophysical Union*, 29, 664-670, 1948.
- 293 Frazer, G. W., Canham, C. D., and Lertzman, K. P.: Gap light analyzer (GLA) version 2.0: imaging
294 software to extract canopy structure and gap light transmission indices from true-colour fisheye
295 photographs, users manual and program documentation. Burnaby (British Columbia (USA) &
296 Millbrook, New York (USA)): Simon Fraser University & The Institute of Ecosystem
297 Studies.Galletal, 2022.
- 298 Deng, J., Fang, S., Fang, X., Jin, Y., Kuang, Y., Lin, F., Liu, J., Ma, J., Nie, Y., Ouyang, S., Ren, J., Tie,
299 L., Tang, S., Tan, X., Wang, X., Fan, Z., Wang, Q.-W., Wang, H., and Liu, C.: Forest understory
300 vegetation study: current status and future trends, *Forestry research*, 3, 6, <https://doi.org/10.48130/FR->
301 2023-0006, 2023.
- 302 Gall, C., Nebel, M., Quandt, D., Scholten, T., and Seitz, S.: Pioneer biocrust communities prevent soil
303 erosion in temperate forests after disturbances, *Biogeosciences*, 19, 3225–3245,
304 <https://doi.org/10.5194/bg-19-3225-2022>, 2022.

305 Geißler, C., Kühn, P., Shi, X., and Scholten, T.: Estimation of throughfall erosivity in a highly diverse
306 forest ecosystem using sand-filled splash cups, *J. Earth Sci.*, 21, 897–900, 2010.

307 Geißler, C., Kühn, P., Böhnke, M., Bruelheide, H., Shi, X., and Scholten, T.: Splash erosion potential
308 under tree canopies in subtropical SE China. *Experiments in Earth surface process research*, 91, 85–
309 93, <https://doi.org/10.1016/j.catena.2010.10.009>, 2012.

310 Goebes, P., Bruelheide, H., Härdtle, W., Kröber, W., Kühn, P., Li, Y., Seitz, S., von Oheimb, G., and
311 Scholten, T.: Species-specific effects on throughfall kinetic energy in subtropical forest plantations are
312 related to leaf traits and tree architecture, *PLoS ONE* 10, e0128084, 2015.

313 Holz, D. J., Williard, K. W., Edwards, P. J., and Schoonover, J. E.: Soil erosion in humid regions: A
314 review, *Journal of Contemporary Water Research & Education*, 154(1), 48-59, 2015.

315 Katayama, A., Oyamada, M., Abe, H., Uemori, K., and Hishi, T.: Soil erosion decreases soil microbial
316 respiration in Japanese beech forests with understory vegetation lost by deer, *J. For. Res.*,
317 <https://doi.org/10.1080/13416979.2023.2235499>, 2023.

318 Kawakami, E., Katayama, A., and Hishi, T.: Effects of declining understory vegetation on leaf litter
319 decomposition in a Japanese cool-temperate forest, *J. For. Res.*, 25, 260–268, 2020.

320 Keim, R. F. and Link, T. E.: Linked spatial variability of throughfall amount and intensity during rainfall
321 in a coniferous forest, *Agric. For. Meteorol.*, 248, 15–21,
322 <https://doi.org/10.1016/j.agrformet.2017.09.006>, 2018.

323 Lal, R.: Soil conservation and ecosystem services, *Int. Soil Water Conserv. Res.*, 2, 36-47, 2014.

324 Levia, D. F., Hudson, S. A., Llorens, P., and Nanko, K.: Throughfall drop size distributions: a review and
325 prospectus for future research, *WIREs Water*, 4, e1225, <https://doi.org/10.1002/wat2.1225>, 2017.

326 Levia, D. F., Nanko, K., Amasaki, H., Giambelluca, T. W., Hotta, N., Iida, S., Mudd, R. G., Nullet, M.
327 A., Sakai, N., Shinohara, Y., Sun, X., Suzuki, M., Tanaka, N., Tantasirin, C., and Yamada, K.:
328 Throughfall partitioning by trees, *Hydrol. Process*, 33, 1698–1708, <https://doi.org/10.1002/hyp.13432>,
329 2019.

330 Liu, J., Liu, W., Zhang, W., and Wang, P.: The detected effects of driptips and leaf wettability on the
331 kinetic energy of throughfall in rubber-based agroforestry in Xishuangbanna, *Ecohydrology*, e2492,
332 <https://doi.org/10.1002/eco.2492>, 2022.

333 Misik, T., and Kárász, I.: Low understory condition in an oak forest in Hungary, 1972 and 2022 –
334 Síkfőkút Project is 50 years old, *Acta Biologica Plantarum Agriensis*, 22–35,
335 <http://doi.org/10.21406/abpa.2022.10.22>, 2022.

336 Miura, S., Yoshinaga, S., and Yamada, T.: Protective effect of floor cover against soil erosion on steep
337 slopes forested with *Chamaecyparis obtusa* (hinoki) and other species, *J. Forest Res.*, 8, 27-35, 2003.

338 Murata, I., Saruki, S., Kubota, K., Inoue, S., Tashiro, N., Enoki, T., Utsumi, and Y. Inoue, S.: Effects of
339 sika deer (*Cervus nippon*) and dwarf bamboo (*Sasamorpha borealis*) on seedling emergence and
340 survival in cool-temperate mixed forests in the Kyushu Mountains, *J. Forest Res.*, 14(5), 296-301,
341 2009.

342 Nanko, K., Mizugaki, S., and Onda, Y.: Estimation of soil splash detachment rates on the forest floor of
343 an unmanaged Japanese cypress plantation based on field measurements of throughfall drop sizes and
344 velocities, *Catena*, 72, 348–361, 2008.

345 Nanko, K., Watanabe, A., Hotta, N., and Suzuki, M.: Physical interpretation of the difference in drop size
346 distributions of leaf drips among tree species, *Agric. For. Meteorol.*, 169, 74–84,
347 <https://doi.org/10.1016/j.agrformet.2012.09.018>, 2013.

348 Nanko, K., Giambelluca, T. W., Sutherland, R. A., Mudd, R. G., Nullet, M. A., and Ziegler, A. D.:
349 Erosion potential under *Miconia calvescens* stands on the island of Hawai'i, *Land Degrad. Dev.* 26,
350 218–226, 2015.

351 Nanko, K., Onda, Y., Kato, H., and Gomi, T.: Immediate change in throughfall spatial distribution and
352 canopy water balance after heavy thinning in a dense mature, *Ecohydrology*, 9(2), 300-314, 2016.

353 Nanko, K., Tanaka, N., Leuchner, M., and Levia, D. F.: Throughfall Erosivity in Relation to Drop Size
354 and Crown Position: A Case Study from a Teak Plantation in Thailand, *Forest-Water Interactions*,
355 *Ecological Studies*. Springer International Publishing, Cham, 279–298, [https://doi.org/10.1007/978-3-](https://doi.org/10.1007/978-3-030-26086-6_12)
356 [030-26086-6_12](https://doi.org/10.1007/978-3-030-26086-6_12), 2020.

357 Nanko, K., Keim, R. F., Hudson, S. A. and Levia, D. F.: Throughfall drop sizes suggest canopy flowpaths
358 vary by phenophase, *Journal of Hydrology*, 612, 128–144, 2022.

359 Pimentel, D. and Burgess, M.: Soil erosion threatens food production, *Agriculture*, 3, 443–463, 2013.

360 R Core Team: R: A Language And Environment for Statistical Computing, R Foundation for Statistical
361 Computing, Vienna, Austria. URL: <https://www.R-project.org/>, 2019.

362 Safari, A., Kavian, A., Parsakhoo, A., Saleh, I., and Jordán, A.: Impact of different parts of skid trails on
363 runoff and soil erosion in the Hyrcanian forest (northern Iran), *Geoderma*, 263, 161–167, 2016.

364 Scholten, T., Geißler, C., Goc, J., Kühn, P. and Wiegand, C.: A new splash cup to measure the kinetic
365 energy of rainfall, *Journal of Plant Nutrition and Soil Science*, 174, 596-601, 2011.

366 Seitz, S., Goebes, P., Song, Z., Bruelheide, H., Härdtle, W., Kühn, P., Li, Y. and Scholten, T.: Tree
367 species and functional traits but not species richness affect interrill erosion processes in young
368 subtropical forests, *Soil*, 2, 49-61, 2016.

369 Seitz, S., Nebel, M., Goebes, P., Käppeler, K., Schmidt, K., Shi, X., Song, Z., Webber, C. L., Weber, B.
370 and Scholten, T.: Bryophyte-dominated biological soil crusts mitigate soil erosion in an early
371 successional Chinese subtropical forest, *Biogeosciences*, 14, 5775-5788, 2017.

372 Shinohara, Y., Ichinose, K., Morimoto, M., Kubota, T. and Nanko, K.: Factors influencing the erosivity
373 indices of raindrops in Japanese cypress plantations, *Catena*, 171, 54-61, 2018.

374 Song, Z., Seitz, S., Li, J., Goebes, P., Schmidt, K., Kühn, P., Shi, X., and Scholten, T.: Tree diversity
375 reduced soil erosion by affecting tree canopy and biological soil crust development in a subtropical
376 forest experiment, *For. Ecol. Manag.* 444, 69–77, 2019.

377 Takatsuki, S.: Effects of sika deer on vegetation in Japan: a review, *Biological Conservation*, 142, 1922-
378 1929, 2009.

379 van Dijk, A. I. J. M., Bruijnzeel, L. A., Rosewell, and C. J.: Rainfall intensitykinetic energy relationships:
380 a critical literature appraisal, *J.Hydrol.*, 261, 1–23, 2002.

381 Wang, P.K., and Pruppacher, H.R.: Acceleration to terminal velocity of cloud and raindrops, *J. Appl.*
382 *Meteorol. Climatol.*, 16, 275–280, 1977.

383 Yao, X., Yu, K., Wang, G., Deng, Y., Lai, Z., Chen, Y., Jiang, Y. and Liu, J.: Effects of soil erosion and
384 reforestation on soil respiration, organic carbon and nitrogen stocks in an eroded area of Southern
385 China, *Sci. Total Environ.*, 683, 98-108, 2019.

386 Zemke, J. J.: Runoff and soil erosion assessment on forest roads using a small scale rainfall simulator,
387 *Hydrology*, 3, 25, <https://doi.org/10.3390/hydrology3030025>, 2016.

388

389



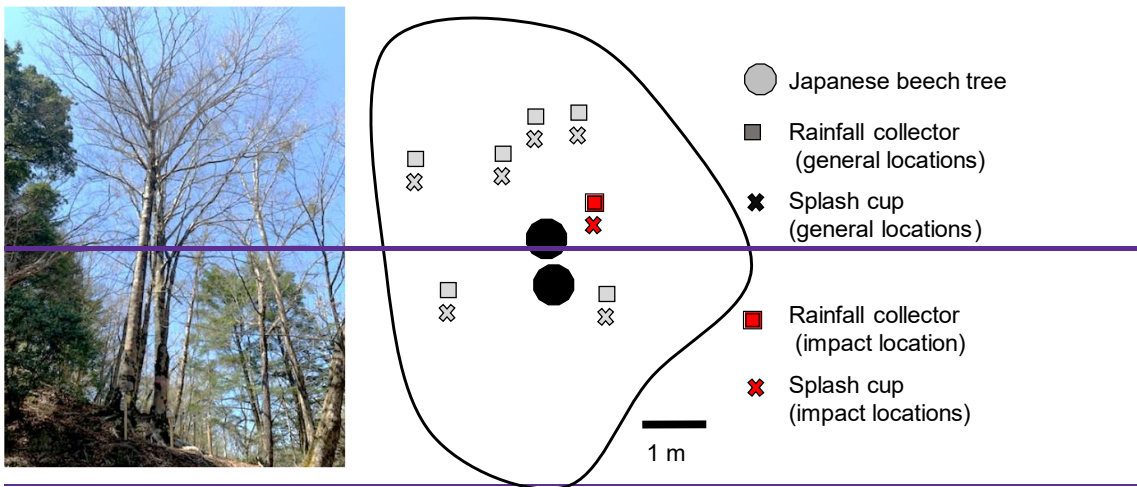
390

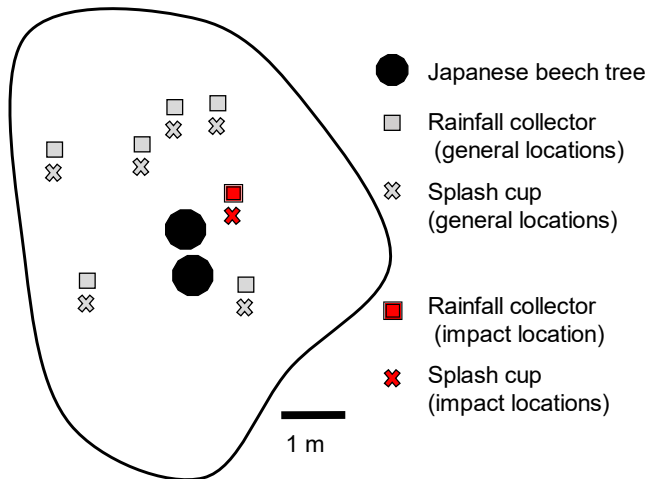
391

392 **Fig. 1** Splash cups at the concentrated impact location (left) and at an exemplary general location (right)

393 after the first rainfall event in the leafless season. Freefall precipitation of this event was 35.4 mm.

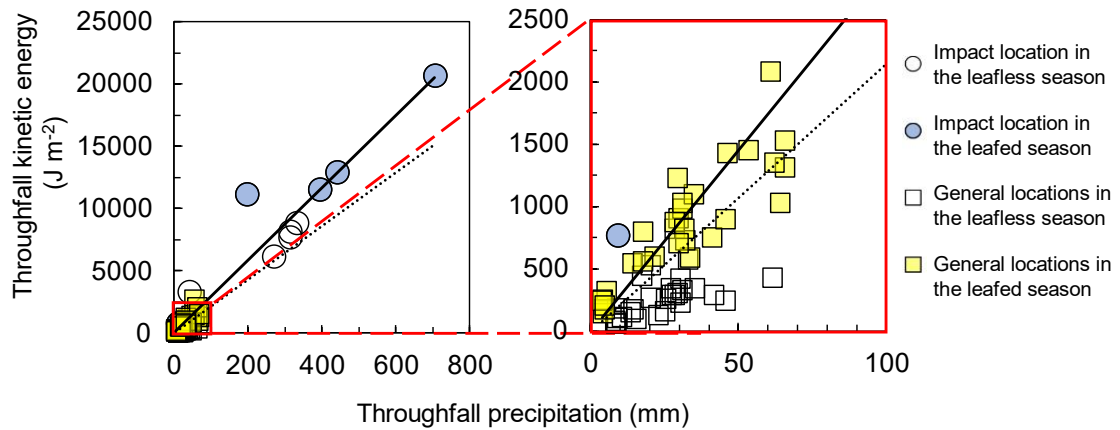
394





396

397 **Fig.2** Japanese beech trees studied in this study (left). The splash cups and rainfall collectors were installed
 398 under the beech trees (right). The black line shows canopy projected area. The splash cups and rainfall
 399 collectors outside the forest were installed 40 m apart from the trees.
 400



401

402 **Fig. 3** Relationship between event-based throughfall precipitation and event-based throughfall kinetic
 403 energy (TKE). Circles and squares show TKE measured at each concentrated impact location and each
 404 TKE at general locations, respectively. Closed and open symbols show leafless and leafed seasons. Solid
 405 and dotted lines show the regression lines at the concentrated impact location and general locations,
 406 respectively. The relationships were significantly different between the locations (ANCOVA, $P < 0.01$).
 407
 408

409

410 **Table 1** Event-scale precipitation, kinetic energy, and unit kinetic energy at the impact location and
411 general locations under Japanese beech trees and outside the forest in the leafless and leafed seasons,
412 respectively.

413

Duration	Precipitation (mm)			Kinetic energy (J m ⁻²)			Unit kinetic energy (J m ⁻² mm ⁻¹)		
	Impact location	General locations	Freefall	Impact location	General locations	Freefall	Impact location	General locations	Freefall
Leafless									
3/3-7	331.7	26.1 ± 8.9	36.0 ± 0.4	8869	274 ± 157	161 ± 20	26.7	11.5 ± 8.5	4.5 ± 0.5
3/11-13	40.4	9.1 ± 0.8	11.9 ± 0.2	3307	102 ± 43	48 ± 2.9	81.9	11.2 ± 4.7	4.0 ± 0.3
3/19-22	314.4	37.1 ± 14.0	43.4 ± 0.7	7737	396 ± 166	385 ± 77	24.6	9.5 ± 2.3	8.9 ± 1.9
3/27-29	314.4	31.0 ± 7.3	38.8 ± 0.7	8166	387 ± 222	294 ± 19	26.0	13.1 ± 8.1	7.6 ± 0.4
4/3-5	268.2	20.5 ± 8.5	24.8 ± 0.2	6182	291 ± 188	25 ± 11	23.1	13.8 ± 6.9	1.0 ± 0.5
Leafed									
8/19-21	445.3 ^a	39.1 ± 12.9	54.1 ± 1.3	11571 ^a	893 ± 189	561 ± 47	26.0	24.2 ± 7.6	10.4 ± 0.9
9/2-3	9.4	4.5 ± 0.5	5.1 ± 0.3	769	223 ± 63	27 ± 8	81.6	49.7 ± 13.7	5.2 ± 1.3
9/10-16	797.5 ^a	56.9 ± 7.3	97.0 ± 1.4	20723 ^a	1723 ± 560	322 ± 50	26.0	30.9 ± 11.4	3.3 ± 0.5
9/27-10/1	498.6 ^a	38.8 ± 14.6	60.6 ± 1.9	12955 ^a	1014 ± 303	7 ± 1.4	26.0	27.4 ± 7.9	0.1 ± 0.0
10/8-11	223.7 ^a	22.0 ± 7.9	27.2 ± 1.5	11137	706 ± 186	12 ± 5.7	49.8	33.3 ± 7.7	0.5 ± 0.2

414 Data are given as mean ± standard deviation.

415 ^a The data was estimated from freefall precipitation.

416